



Analysis of photocatalytic degradation of azo dyes under sunlight with response surface method

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ABSTRACT

After synthesizing evaporating induction by ammonia to build dopant, the response surface method and central composite design model were used to find the optimum conditions of Direct Blue 71 dye removal. This approach was also used to determine interaction design parameters (pH, dye concentration, Zn:Ti molar ratio, photocatalyst concentration and time). Results showed that TiO₂ doped activity was higher than that of TiO₂ under sunlight. Additionally, the obtained model shows a positive correlation between the experiment results and the predicted response surface method responses. The removal efficiency was found to be 87% at 6 h and 90% at 8 h in optimal dye removal conditions including a pH of 6.9, a photocatalyst concentration of 44.6 mg/m², a dye concentration of 53 mg/L, and a Zn:Ti of 0.53%.

Keywords: Central composite design; Zn-TiO₂; DB71; Optimization; photo-reactors

1. Introduction

Dyes play important roles in various parts of the textile industry. Dyes often are artificial and obtained from two sources of oil: intermediate level and coal tar [1]. Synthetic dyes have a high solubility in water. Approximately 15% of the world's dyes enter industrial wastewaters and the environment during the dyeing process. In addition to the textile industry, the leather, paper, food, and health industries, as well as light absorbing cells and electrochemical photovoltaic cells also have dye in their wastewaters [2].

Heterogeneous photocatalyst by semiconductors is a common treatment method derived from advanced oxidation types that shows great potential for monitoring organic contaminants in water [3]. In the past decade, numerous studies conducted by researchers around the world have examined the use of heterogeneous photocatalyst oxidation in the decomposition and mineralization process of particular pollutants [2]. Studies have also looked at light irradiation's effect on the photocatalyst activity

of different forms of TiO₂ such as film [4], powder [5], nanotube [6], stabilized [7], and doped TiO₂ [8].

Stimulation of TiO₂ layers by UV light leads to excitement in the electron capacity layer by the absorption of the photon and its transport layer in the direction of stimulation. In this case, electron-hole pairs are generated at the nanoparticle level, and oxygen molecules hitting the surface capture these electrons. The material level is thus very active, so the water can be oxidized by it. Therefore, this way of dealing with organic carbon molecules could transform them into carbon dioxide, water, and other formats [9].

One of the main problems of using TiO₂ as catalyst is its low efficiency in the visible light spectrum, which is due to a quick remix of electrons and holes. TiO₂ is not active due to its large band gap in visible light, so it is not able to maximize its solar photocatalyst potential. Many methods have been applied to TiO₂ to help it attract visible low-energy photons; these methods include modifying the surface by organic materials, grafting with semi-conductors, and modifying the band gap. To modify the band gap and improve efficiency, and reduce electronic remix, transition metals can be added as a dopant to TiO₂ [10–12].

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Due to the overlap of the intermediate band (3d) titanium with an orbital of d level in the transition metals, using a suitable transition metal as a dopant improves TiO₂ performance [13]. In general, making dope with the TiO₂ and metal ion improves the potential in generated radicals and broadens the absorption spectrum, making it less likely that electrons will become inhibited by ions, other electrons, and the hole remix [2].

Many researchers have investigated the use of dopant for increasing the photocatalyst activity of TiO₂. For example, Narayana et al. [14] investigated the photocatalyst effect of Fe and Co doped with TiO₂ using dye removal at 100 mg/L green malachite with 0.1 g catalyst in pH 7.3 ~ 7.5 in the sun. The rate of dye removal with Fe-TiO₂ after 3 h was more than 98% and 86% with the Co-TiO₂. The kinetics of dye removal with the Langmuir–Hinshelwood model is justifiable based on study results. Reaction rates of dye removal remained constant with Fe-TiO₂ at 0.8 h⁻¹, using dopant Co-TiO₂ at 0.67 h⁻¹, and using pure TiO₂ at 0.31 h⁻¹. Therefore, the photocatalyst effect of Fe-TiO₂ is more than that of the other two catalysts.

Goswami and Ganguli [15] compared the effect of Fe-TiO₂ made with ferrous sulfate hepta hydrate and ferric nitrate nanohydrate. In their study, a 25-W mercury lamp served as a light source. The initial concentration of the dye solution 10–5 was considered; 2 g catalyst in 50 mL dye solution was mixed in; and aeration was conducted. The results show that dopant made with ferrous due to a further reduction of crystalline network TiO₂ has a better photocatalyst effect on the amount of dye removal within 80 min, while the amount of dye removal with the ferric dopant was 80%.

The amount of Congo red dye removal was also tested to examine the photocatalyst effect of titanium dioxide-sulfonic acid nanocomposite irradiated with visible sunlight. In this case, the amount of dye removed from the solution in a volume of 100 mL with a concentration of 40 mg/L, a catalyst of TiO₂/SA, 1.5 g/L, and neutral pH for 210 min was more than 90%. Furthermore, the H₂O₂ oxidizing effect on dye removal was studied using the previous terms but adding 20 mm hydrogen peroxide. The rate of dye removal in 100 min under the same conditions was also more than 90% [16].

Yousef et al. [17] revealed satisfactory photodegradation of azo dyes, reactive black5 (RB5), and methyl red (MR) after being subjected to photocatalytic degradation for 120 min when the temperature was between 24°C and 30°C. Twenty-five mg of nanocatalysts (Cu₀ NPs-decorated, carbon-doped TiO₂ nanofibers, and Cu-free carbon-doped TiO₂ nanofibers) were added to 50 ml (10 mg/l) of the prepared azo dyes. These aqueous solutions were magnetically stirred under sunlight radiation. The sample dye concentration was measured by an UV–vis spectrophotometer. The prepared nanofibers removed 83% from RB5 and 65% from MR after 2 h of reaction.

The photoactivity sensitivity of Cu-TiO₂/ZnO toward operating conditions in an advanced oxidation reaction for the photodegradation of two different dyes under visible light irradiation has been investigated. The Cu-TiO₂/ZnO photocatalyst presented a maximum removal efficiency of color (83.35%), chemical oxygen demand (COD) (73.54%), and total organic carbon (TOC) (54.46%) for methyl orange (MO) in degradation with a concentration of 20 mg/L, 0.7 mg catalyst, and pH 4 for 120 min. For methylene blue (MB) degradation, the maximum removal efficiency was color (75.50%),

COD (68.00%), and TOC (46.41%) with a concentration of 20 mg/L, 0.7 mg catalyst, and pH 10 for 120 min [18].

There are many methods for making dopant including the sol-gel, ball milling, and flame methods. In this study, a dopant with different amounts of metal created by evaporation-induced synthesis was constructed. Concrete was chosen as a stabilizing platform because it is one of the most useful and important constitutive materials of treatment plants. It is also available, inexpensive, and it has a simple construction technology. The experimental design was performed by response surface method (RSM) in order to data analyze and optimize initial pH, photocatalyst concentration, pollutant concentration, time, and ratio of zinc to titanium in the dopant manufacturing stage in synthetic dye removal Direct Blue 71 (DB71) with eight aromatic rings and three benzene rings.

2. Materials and method

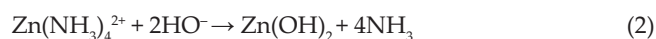
2.1. Materials

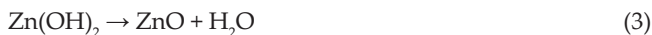
Materials used in this study include DB71 dye made from Alvan Sabet of Hamedan; titanium dioxide nanoparticles containing anatase to rutile with a ratio of 80 to 20 and a 21 nm diameter from US Research NANO Inc. of USA; ammonia from Merck Co. of Germany; Portland concrete and water to produce concrete; concrete epoxy super glue made from optimal concrete chemistry; mold oil; and distilled water.

Equipment used included a Cary 50 spectrophotometer Varian for measuring absorption; Fungilab UE-6SFD ultrasonic cleaners for dispersing nanoparticles of titanium dioxide; Metrohm 691 digital pH meter to determine the pH of samples; Kern PLS 360-3 digital scale; Technic AP 1400F aquarium water pump capable of transferring water to a height of 1.3 m and having a flow rate of 200 L/s; Atinan AP aeration pump to provide required oxygen for the photocatalyst process with a liter per minute aeration flow; TES 1333 solar meter to measure the intensity of solar radiation; and DEM-RED oven with heating capacity to 800°C with the ability to stabilize the nanoparticles.

2.2. Preparation of photocatalyst

In order to make nanocomposite by induction, the evaporating synthesis method with ammonia was used [19]. First, the ammonia was added to nano-TiO₂ to reach a pH of 10.5 and stirred, then the suspension was added to zinc solution to form a complex of zinc ammonium with two additional positive charge (white layer formation on nanocrystal of TiO₂ represents the desired complex formation). This solution was stirred for 24 h according to Eq. (1). In order to convert to hydroxide Zn (Eq. (2)), the solution was evaporated. Finally, to reach the thermal decomposition zinc hydroxide to zinc oxide (Eq. (3)), the nanocomposite was placed in the oven for 4 h at 450°C.





X-ray diffraction (XRD) analysis was used to determine the presence of crystalline phases of titanium dioxide in the dopant composite structure.

2.3. Photochemical reactor start-up

The photo reactor shown in Fig. 1 consists of several parts. The reactor core, in which the solar photocatalytic process was carried out, consists of a 40 cm base of stairs, 30 cm height, and 37° horizontal angles. It is made with metal cans and a corner stone grade 3 for the maintenance of five concrete stairs. The 24 × 12 × 4 concrete plates are made of cement, on which the stabilized nanoparticles were placed during photocatalyst. The plates are 2.5 cm apart, and the ends of each stair have a 4 cm overlap. The 5-L wastewater storage tank is made of galvanized reflected sunlight. To protect the reactor against wind, rain, and evaporation, ordinary 3 mm thick glass of 40 × 100 cm and a 37° horizontal angle at the front of reactor is installed, along with two 40 × 80 cm plates at the sides and two 60 × 40 cm plates behind the reactor. The ordinary glass light transmission coefficient is for wavelengths longer than 350 nm, which is more than 90% or Pyrex. Pyrex was not used, however, due to inaccessibility of the required dimensions among domestic producers and high cost from foreign countries.

To achieve laminar flow, the maximum flow rate is 22.95 L/min. For this reason, a 13.5-W aquarium pump with a flow rate of 200 L/h and a pumping height of 1.3 m is installed. Air is blown through an aerator pump for the dissolved oxygen supply in water. All the mentioned parts are placed on the 40 × 80 × 120 cm chassis. Third grade canisters equipped with four wheels are installed for ease of movement. This study established the solar reactor at the geographical coordinates 510 22' 95" E and 350 43' 12" N and 1,776 m above sea level. The photo reactor under a 37° angle is oriented to the south and contains a solar meter to measure the intensity of solar radiation in the visible region. Among the methods of nanoparticle stabilization, including sol-gel, slurry, and mix, slurry was used due to improved efficiency and ease of implementation [13].

In supplying a sustainability photocatalyst on the concrete context, the durability of the consolidation method is important. In other words, the most important parameter

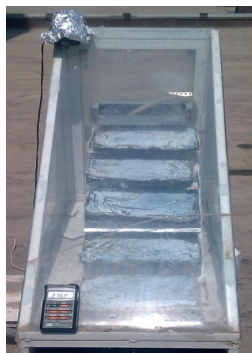
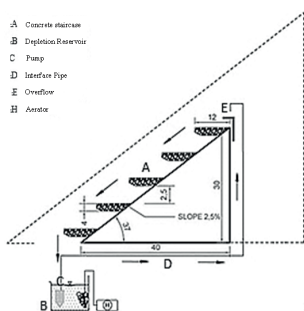


Fig. 1. Applied photo-reactor.

in the coated photocatalytic reactor is the lack of decreasing efficiency of the system during continuous operation due to the separation of the particles from the substrate surface [20].

First, the glue was applied with a paint brush to a thickness of 5 mm on a concrete surface, and the TiO₂ suspension in a solution of distilled water-ethanol was produced. The solution was stirred for 10 min at 1,500 rpm and 200°C temperature. It remained in the ultrasonic cleaner for 20 min to separate aggregate nanoparticles and prepare a homogeneous suspension. After creating a homogeneous suspension, and 30 min after applying the glue, the concrete surface with suspension was submerged. It was dried for 72 h at room temperature in order to increase resistance against platform abrasion, and then heated in the oven at 1,500°C for 8 h.

2.4. The photocatalytic process

All tests were performed between 8 a.m. and 4 p.m. under sunlight, moderated at 1,004 W/m² and an average temperature of 35°C. Factors examined in this study included pH (6.9, [natural pH of dyes], 9, and 11), different molar ratios Zn:Ti (0.05%, 0.33%, 0.53%, 0.72%, and 1%), dye concentration (5, 20, 50, and 100 mg) and photocatalyst concentration (20, 50, and 80 g/m²). To perform the tests, 3 L of synthetic wastewater (DB71 combination of dye and distilled water) was used. The percentage removal of dye at a rate of absorption at the wavelength of maximum absorbance (586 nm) using a spectrophotometer was determined.

3. Result

3.1. Characterization of Zn-TiO₂

Fig. 2 presents the XRD analysis of titanium dioxide nanoparticles. When comparing the samples' intensity of the characteristic peak of anatase (25.4, I_A) and rutile (27.4, I_R), the ratio of anatase to rutile phase (I_R/I_A) in doped samples appears to be fairly equal to the undoped sample, meaning the doping process inhibits anatase transmission to rutile [21]. It should be noted that the photocatalytic properties of anatase is superior to those of the rutile phase [22].

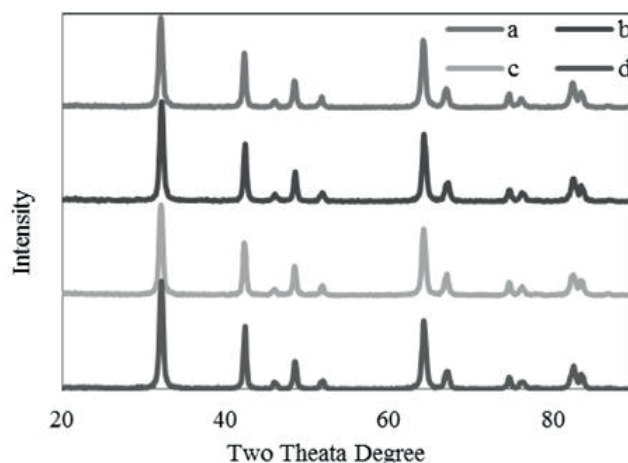


Fig. 2. XRD of: a – undoped TiO₂, b – 0.5% Zn-TiO₂, c – 0.6% Zn-TiO₂, and d – 0.72% Zn-TiO₂.

3.2. RSM model analysis

3.2.1. Design

Table 1 presents the central composite design [23] table, along with the experimental and predicted responses. Point 38 in the model, removal of dye, and point 1, removal of organic load, were ignored. A quadratic equation was obtained for all three responses using the analysis of variance results with a 95% confidence level. Table 2 reports the standard errors, coefficients of reliability, sum of squares, and *F*-value. Eqs. (1)–(3), respectively, outline the proposed model for dye removal, removal of organic load, and the pH value of the output. The model related to the removal of organic load by removing two sentences $[\text{Dye}]^2$ and $[\text{time}]^2$ has been modified. Values of adjusted R^2 and R^2 are reported in Table 2.

$$\begin{aligned} \text{Dye removed \%} = & -332.93216 - 0.74502 \times [\text{Dye}] + 1.86425 \\ & \times \text{pH} + 9.75641 \times [\text{ZnO-TiO}_2] + 37.74798 \times \text{Time} + 499.16242 \\ & \times [\text{Zn:Ti}] - 1.14883\text{E-}003 \times [\text{Dye}] \times \text{pH} + 0.010624 \times [\text{Dye}] \\ & \times [\text{ZnO-TiO}_2] + 0.022204 \times [\text{Dye}] \times \text{Time} - 0.40872 \times [\text{Dye}] \\ & \times [\text{Zn:Ti}] + 2.92746\text{E-}003 \times \text{pH} \times [\text{ZnO-TiO}_2] - 0.052473 \times \text{pH} \\ & \times \text{Time} - 1.27678 \times \text{pH} \times [\text{Zn:Ti}] - 0.065152 \times [\text{ZnO-TiO}_2] \\ & \times \text{Time} - 4.17371 \times [\text{ZnO-TiO}_2] \times [\text{Zn:Ti}] - 9.85275 \times \text{Time} \\ & \times [\text{Zn:Ti}] - 9.61007\text{E-}004 \times [\text{Dye}]^2 - 0.20428 \times \text{pH}^2 - 0.091253 \\ & \times [\text{ZnO-TiO}_2]^2 - 2.56921 \times \text{Time}^2 - 203.14123 \times [\text{Zn:Ti}]^2 \quad (4) \end{aligned}$$

$$\begin{aligned} \text{COD removed \%} = & -41.21921 + 0.38770 [\text{Dye}] + 19.84771 \text{pH} \\ & + 2.06753 [\text{ZnO-TiO}_2] - 3.61193 \text{Time} - 44.83242 [\text{Zn:Ti}] \\ & - 0.069399 \times [\text{Dye}] \times \text{pH} - 1.35125\text{E-}003 \times [\text{Dye}] \times [\text{ZnO-} \\ & \text{TiO}_2] + 0.031382 \times \text{Dye} \times \text{Time} + 0.82185 \times [\text{Dye}] \times [\text{Zn:} \\ & \text{Ti}] - 0.028452 \times \text{pH} \times [\text{ZnO-TiO}_2] - 0.63157 \times \text{pH} \times \text{Time} \\ & + 9.93383 \times \text{pH} \times [\text{Zn:Ti}] + 0.098725 \times [\text{ZnO-TiO}_2] \times \text{Time} \\ & - 0.43868 \times [\text{ZnO-TiO}_2] \times [\text{Zn:Ti}] + 1.21132 \times \text{Time} \times [\text{Zn:} \\ & \text{Ti}] - 0.96218 \times \text{pH}^2 - 0.019572 \times [\text{ZnO-TiO}_2]^2 - 71.51145 \\ & \times [\text{Zn:Ti}]^2 \quad (5) \end{aligned}$$

$$\begin{aligned} \text{pH-out} = & +8.00944 + 0.016456 \times [\text{Dye}] - 0.26458 \times \text{pH} - \\ & 3.71064\text{E-}003 \times [\text{ZnO-TiO}_2] + 0.092280 \times \text{Time} + 1.52993 \\ & \times [\text{Zn:Ti}] - 5.95458\text{E-}004 \times [\text{Dye}] \times \text{pH} - 2.25778\text{E-}004 \\ & \times [\text{Dye}] \times [\text{ZnO-TiO}_2] + 1.91397\text{E-}004 \times [\text{Dye}] \times \text{Time} - \\ & 6.26798\text{E-}004 \times [\text{Dye}] \times [\text{Zn:Ti}] + 1.35529\text{E-}003 \times \text{pH} \times [\text{ZnO-} \\ & \text{TiO}_2] - 5.05076\text{E-}003 \times \text{pH} \times \text{Time} - 0.27726 \times \text{pH} \times [\text{Zn:Ti}] \\ & + 6.39763\text{E-}004 \times [\text{ZnO-TiO}_2] \times \text{Time} - 0.019849 \times [\text{ZnO-} \\ & \text{TiO}_2] \times [\text{Zn:Ti}] - 0.038279 \times \text{Time} \times [\text{Zn:Ti}] - 8.75158\text{E-}006 \\ & \times [\text{Dye}]^2 + 0.030953 \times \text{pH}^2 + 2.72505\text{E-}004 \times [\text{ZnO-TiO}_2]^2 \\ & - 6.10169\text{E-}003 \times \text{Time}^2 + 2.08423 \times [\text{Zn:Ti}]^2 \quad (6) \end{aligned}$$

Fig. 3 presents the degree of dispersion of experimental data and predicted responses (indicating the R^2 coefficient). The dispersion of points around and below the 45° line indicates that the model predicted values are closer to reality and are more accommodating [24].

To use the model in the design space, the ratio of S:N (signal to noise) must be greater than 4 to achieve the model of dye removal of 15.83, the removal of organic load of 7.224, and the pH output of 6.661.

3.2.2. Analysis of effective factors

After introducing the models and assessing their errors, the interaction of effective factors was investigated. Fig. 4 shows the effect of dye concentration and pH on the responses of dye removal, removal of organic load, and output pH in constant conditions of concentration: photocatalyst of 40 g/m², the Zn:Ti ratio equal to 0.5, pH 7, and duration of 5 h. Fig. 4(A) shows that, with increasing dye concentration and pH, the organic load removal efficiency of less than 36 increases to more than 90% because the complex building dye decreases. This results in an increased removal of organic load. Fig. 4(B) indicates that an increase in the acidic pH of the dye concentration increases output pH, and the alkaline pH leads to decreased output pH because the dye investigated has been anion and is buffered.

According to Fig. 4(C), as pH increases, dye removal decreases due to the decreasing complex building process. As dye is increased, efficiency decreases because there is insufficient reaction time to remove the dye [25,26].

Fig. 5 shows the effect of dye concentration and a photocatalyst on the responses' dye removal, and output pH during constant conditions (Zn:Ti ratio equal to 0.5, pH 7, and 5-h duration). Fig. 5(A) shows that the amount of photocatalyst in the dye removal reduces as dye concentration increases. This reduces the effect of the surface's number of active centers relative to the contaminant molecules. As Fig. 5(B) indicates, by increasing the amount of photocatalyst, output pH increases because the doped metal produces hydroxyl and oxygen radicals [27].

Fig. 6 shows the effect of dye concentration and the Zn:Ti molar ratio on dye removal responses and removal of organic load in constant conditions (photocatalyst concentration 40 g/m², pH 7, and 5-h duration). In Fig. 6(A), the increase in dye concentration decreases the effect of the Zn:Ti molar ratio in the removal efficiency because the number of hydroxyl radicals generated by doped metal rather than the pollutant molecules is decreased. In the composite TiO₂-ZnO, the electron transfers from the conduction band of excited ZnO to the conduction band of excited TiO₂ and vice versa; a cavity transfer can occur from the valence band ZnO to the valence band TiO₂; this effective separation increases the responsibility load of the composite photocatalytic. Therefore, the maximum photocatalytic activity was related to the amount of ZnO, which received electrons generated from the TiO₂ light excitation. Therefore, as Fig. 6(B) shows, when the optimum photocatalytic activity is approached, more electrons are generated, and the output pH decreases. Jiang et al. [28] reported similar results in their study.

Fig. 7 demonstrates the effect of the photocatalyst concentration and pH on dye removal responses and removal of organic load in constant conditions of dye concentration 40 mg/L, the Zn:Ti molar ratio equal to 0.5, and 5-h duration. Figs. 7(A) and (B) show that by increasing pH, dye removal efficiency decreases, and the organic load removal efficiency increases because the complex process decreases. The dye removal efficiency of the organic load increases with an increased concentration of photocatalyst, but after about 50 g/m², it is decreased since the increased photocatalyst leads to increased active surface sites, thereby increasing the

Table 1
CCD matrix

Factor No.	Dye con. (mg/L)	pH	Photo catalysis (g/m ²)	Time (h)	Zn:Ti	Experimental			RSM		
						DR ^a %	CODR ^b %	pH-out	DR %	CODR %	pH-out
1	33	5	37	3	0.33	54.6	59.2	8.3	55.1	–	8.3
2	72	5	37	3	0.33	29.1	88.4	8.6	34.3	86.3	8.5
3	33	9	37	3	0.33	42.5	92.4	8.6	50.0	87.8	8.6
4	72	9	37	3	0.33	27.0	91.1	8.8	29.1	91.7	8.7
5	33	5	63	3	0.33	39.1	85.3	9.1	40.9	74.9	8.8
6	72	5	63	3	0.33	31.7	86.7	8.8	30.8	86.8	8.7
7	33	9	63	3	0.33	44.2	72.2	9.4	36.1	87.2	9.2
8	72	9	63	3	0.33	19.5	96.5	9.1	25.8	89.7	9.1
9	33	5	37	6	0.33	77.6	42.5	8.4	82.8	67.6	8.4
10	72	5	37	6	0.33	61.8	93.3	8.5	64.6	84.5	8.6
11	33	9	37	6	0.33	73.4	88.9	8.8	77.3	76.1	8.7
12	72	9	37	6	0.33	52.0	6.1	9.2	58.9	83.7	8.8
13	33	5	63	6	0.33	57.8	96.0	9.2	63.8	76.8	8.9
14	72	5	63	6	0.33	52.6	75.3	8.9	56.3	92.3	8.9
15	33	9	63	6	0.33	63.6	65.1	9.3	58.5	82.8	9.3
16	72	9	63	6	0.33	42.6	98.2	9.0	50.8	89.0	9.2
17	33	5	37	3	0.72	87.3	56.8	9.2	87.0	51.9	8.9
18	72	5	37	3	0.72	48.5	69.3	9.2	59.7	78.2	9.0
19	33	9	37	3	0.72	80.7	78.2	9.1	80.2	80.0	8.8
20	72	9	37	3	0.72	49.6	90.5	8.8	52.8	97.0	8.9
21	33	5	63	3	0.72	33.0	23.5	9.0	30.7	49.3	9.1
22	72	5	63	3	0.72	10.2	81.3	9.1	14.1	74.2	9.1
23	33	9	63	3	0.72	16.5	78.2	9.4	24.2	74.9	9.2
24	72	9	63	3	0.72	9.2	91.1	9.4	7.5	90.6	9.1
25	33	5	37	6	0.72	95.1	46.1	9.1	103.1	47.8	8.9
26	72	5	37	6	0.72	72.3	81.8	9.1	78.4	77.8	9.1
27	33	9	37	6	0.72	94.2	50.8	8.8	95.9	69.7	8.8
28	72	9	37	6	0.72	59.5	86.2	8.9	71.0	90.3	8.9
29	33	5	63	6	0.72	44.8	41.3	9.3	42.1	52.6	9.2
30	72	5	63	6	0.72	20.9	75.8	9.4	28.1	81.2	9.2
31	33	9	63	6	0.72	28.3	77.0	9.4	35.0	72.0	9.2
32	72	9	63	6	0.72	20.9	92.2	9.3	20.9	91.3	9.1
33	5	7	50	5	0.53	98.9	81.4	8.3	99.2	67.8	8.7
34	100	7	50	5	0.53	76.6	91.6	8.6	57.6	106.4	8.7
35	53	3	50	5	0.53	95.0	70.1	8.8	84.6	57.0	9.1
36	53	11	50	5	0.53	78.3	86.4	9.1	70.0	86.4	9.4
37	53	7	20	5	0.53	55.4	77.5	8.3	36.7	67.9	8.6
38	53	7	80	5	0.53	47.2	74.6	9.1		71.1	9.4
39	53	7	50	1	0.53	26.7	95.3	8.2	24.6	89.9	8.6
40	53	7	50	8	0.53	90.2	93.1	8.6	73.6	84.3	8.7

(continued)

Table 1 (continued)

Factor No.	Dye con. (mg/L)	pH	Photo catalysis (g/m ²)	Time (h)	Zn:Ti	Experimental			RSM		
						DR ^a %	CODR ^b %	pH-out	DR %	CODR %	pH-out
41	53	7	50	5	0.05	40.0	79.8	8.9	33.6	82.2	8.9
42	53	7	50	5	1.00	48.2	75.3	9.0	35.9	59.7	9.5
43	53	7	50	5	0.53	84.5	93.1	8.3	80.6	87.1	8.7
44	53	7	50	5	0.53	81.5	88.6	8.7	80.6	87.1	8.7
45	53	7	50	5	0.53	81.3	88.6	8.8	80.6	87.1	8.7
46	53	7	50	5	0.53	82.2	89.4	8.7	80.6	87.1	8.7
47	53	7	50	5	0.53	82.0	89.4	8.7	80.6	87.1	8.7
48	53	7	50	5	0.53	82.3	90.1	8.7	80.6	87.1	8.7
49	53	7	50	5	0.53	81.4	88.6	8.8	80.6	87.1	8.7
50	53	7	50	5	0.53	81.6	90.9	8.8	80.6	87.1	8.7

^aDR: Dye removal.^bCODR: The expression of the removal of the organic matter.Table 2
Analysis of variance

Response	Source	Squares	df	Square	F-value	Prob > F	
DR	Model	29,475.11	20	1,473.756	17.28648	<0.0001	Significant
	Residual	2,387.136	28	85.25484			
	Lack of fit	2,379.479	21	113.3085	103.5942	<0.0001	Significant
	Pure error	7.656409	7	1.093773		R ² value = 0.9251	
	Cor. total	31,862.25	48			R ² adj = 0.8716	
CODR	Model	7,906.20	18	439.23	2.59	0.0103	Significant
	Residual	5,088.25	30	169.61			
	Lack of fit	5,071.9	23	220.52	94.37	<0.0001	Significant
	Pure error	16.36	7	2.34		R ² value = 0.6084	
	Cor. total	12,994.45	48			R ² adj = 0.37735	
pH-out	Model	3.454548	20	0.172727	2.451472	0.0136	Significant
	Residual	2.0433	29	0.070459			
	Lack of fit	1.8833	22	0.085605	3.7452	0.0396	Significant
	Pure error	0.16	7	0.022857		R ² value = 0.683	
	Cor. total	5.497848	49			R ² adj = 0.3720	

removal of dyes and organic load. However, if the amount of photocatalyst is much more efficient, photocatalyst particles agglomerate, and the number of active points decreases because light does not reach the compacted particles. This results in decreases in dye removal efficiency and organic load; Sajjad et al. [29] and Sun et al. [30] report similar results.

The interaction of two factors, pH and Zn:Ti, during a 5-h duration, photocatalyst concentration of 40 g/m², and dye concentration of 40 mg/L, is shown in Fig. 8, as well as its impact on dye removal, organic load removal, and output pH. Zinc ion as a dopant produces oxygen and hydroxyl radicals, which effectively reduces the dye amount [13]. Thus, when increasing the amount of zinc ion on dye removal, removal of organic load increases because the concentration

of the doped TiO₂ is higher than optimal, the oxygen surface forms cavities, and Ti³⁺ acts as a recombination center to decrease the doped samples' activity in visible light [11] (Figs. 8(A) and (B)). In Part C, the increase of Zn:Ti due to additional hydroxyl radical production causes pH output to increase.

In Fig. 9, the effect of the interaction between the dye concentration and duration on dye removal and removal of organic dyes has been shown, with a photocatalyst concentration of 40 g/m², pH 7, and Zn:Ti ratio equal to 0.5. The effect of the photocatalyst concentration on dye removal increases over time (Fig. 9(A)) and decreases the organic load degradation (Fig. 9(B)). The increase of time enables sufficient time for a reaction, but the production power of the radicals in a

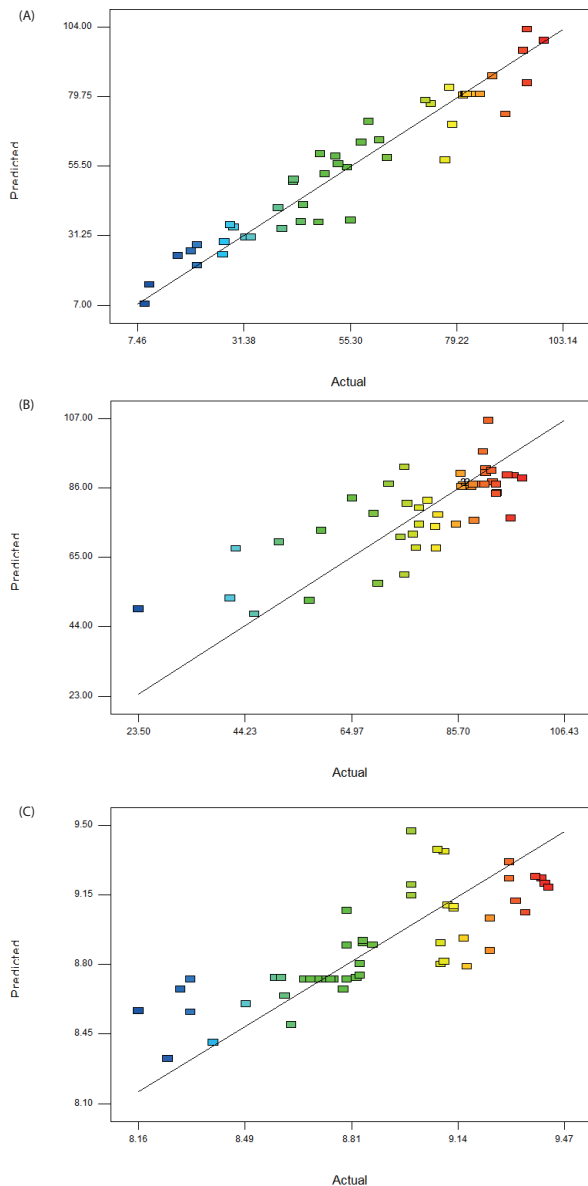


Fig. 3. The experimental response against the predicted values: (A) DR, (B) CODR, and (C) pH-out.

certain amount of metal (Zn:Ti fixed) does not increase with time; therefore, organic load removal will reduce [31].

Fig. 10 shows the interaction between the photocatalyst concentration and the Zn:Ti ratio on dye removal in constant conditions (5 h, 40 mg/L dye concentration, and neutral pH), indicating the reduction of the optimum photocatalyst amount with increasing Zn:Ti, because the metal doped photocatalytic activity exacerbates TiO_2 due to increased hydroxyl radical production [32].

Fig. 11 provides two-dimensional diagrams representing the effect of time and Zn:Ti on dye removal responses and output pH in constant conditions (photocatalyst concentration of 40 g/m² and dye concentration of 40 mg/L at neutral pH). By increasing time, the effect of the Zn:Ti ratio on dye removal increases because there is ample opportunity for response and increased radical production. In this case,

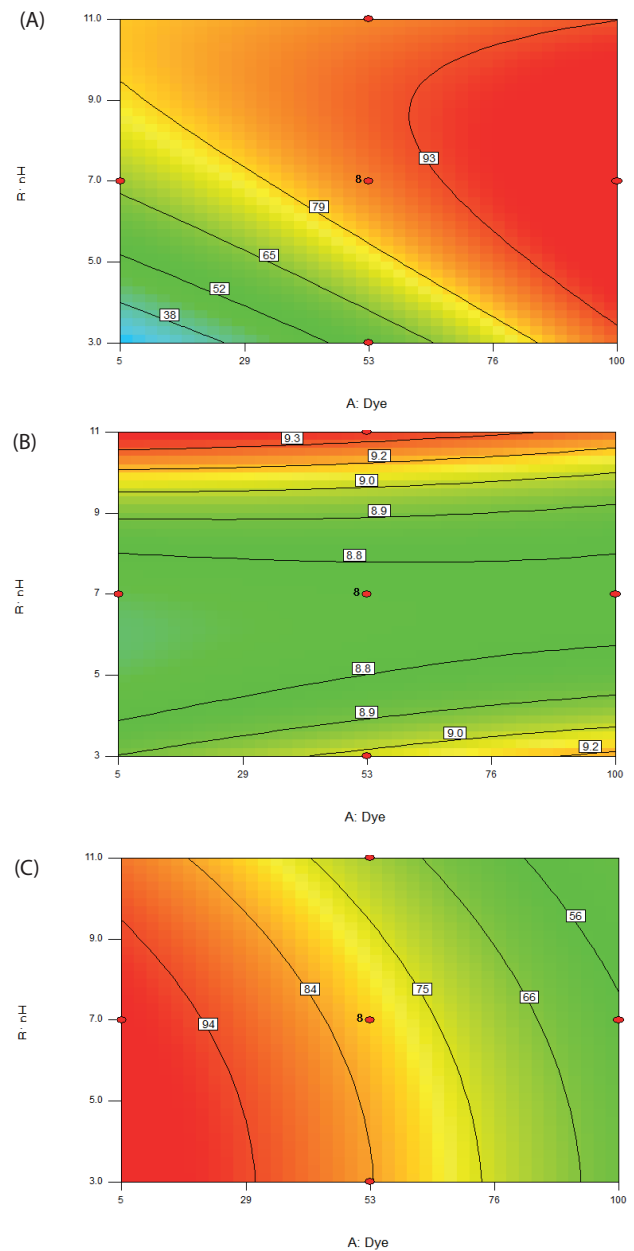


Fig. 4. Effect of initial dye concentration and pH: (A) CODR, (B) pH-out, and (C) DR ($t = 5$ h, Zn:Ti = 0.5%, $[\text{Zn-TiO}_2] = 40$ g/m²).

output pH increases, similar to the results found by Nguyen et al. [32] and Rehman et al. [11].

3.3. The photocatalytic degradation of dye

3.3.1. Effect of pH

Changes in pH solution changes TiO_2 particle surface charge and, consequently, catalytic reaction [33]. Fig. 12 outlines the results of pH changes at constant conditions. According to the results, the system's efficiency at pH 3 to 11 ranges from 90% to 40%, so a complete model is needed in all these areas. The TiO_2 surface below the ISO-electric point (6.8) is positive but negative above the ISO-electric point.

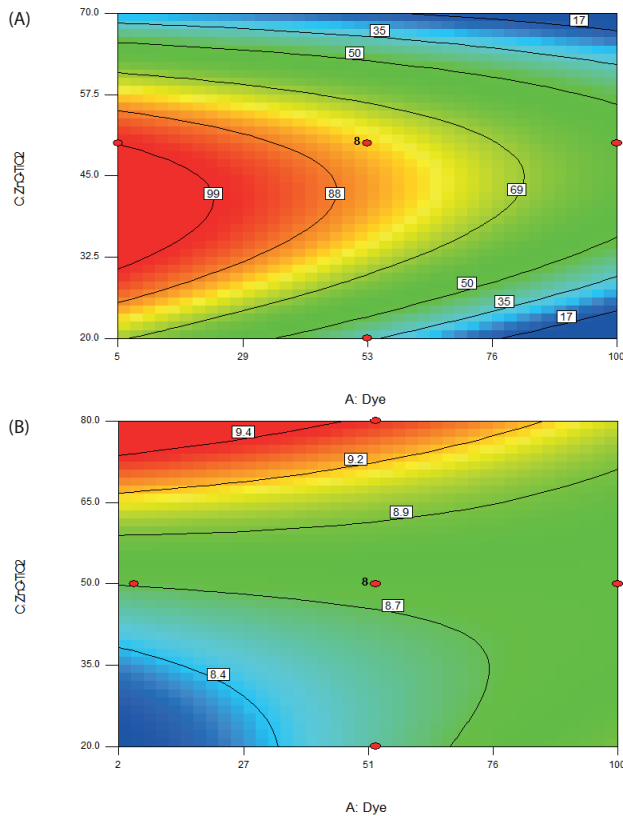


Fig. 5. Effect of initial dye concentration and amount of photocatalysis: (A) DR and (B) pH-out ($t = 5$ h, Zn:Ti = 0.5%, pH = 7).

Therefore, the surface adsorption of a given dye structure may be more or less in acidic or basic environments, but in general, in the ISO-electric point (point with zero charge), it is minimized [33].

As shown in Fig. 13, anionic dyes investigated in acidic conditions clot [25], and coagulated grains of dye sit on the catalyst surface and cause the photocatalyst operation to malfunction. Even when the efficiency is high, the unremoved dye is deposited on the substrate.

3.3.2. Effect of dye concentration

Fig. 14 indicates the effect of the concentration in inconstant conditions with increasing inlet pollutant concentrations ranging from 10 to 100 mg/L in a fixed system. The efficiency decreases from 98% to 50% since the effect of light decreases with increasing dye concentration and the number of active centers on the photocatalyst reduce. Chong et al. [33] and Jiang et al. [28] report similar results in their respective studies.

3.3.3. Effect of photocatalyst concentration

According to Fig. 15, by increasing the photocatalyst amount from 20 to 42 g/m², the dye removal increases from 68% to 90% because the number of active centers on the surface and the absorption amount of photons and pollutants molecules on the catalyst surface increase. However, after increasing the photocatalysts over 42 g/m², efficiency decreases because agglomerate

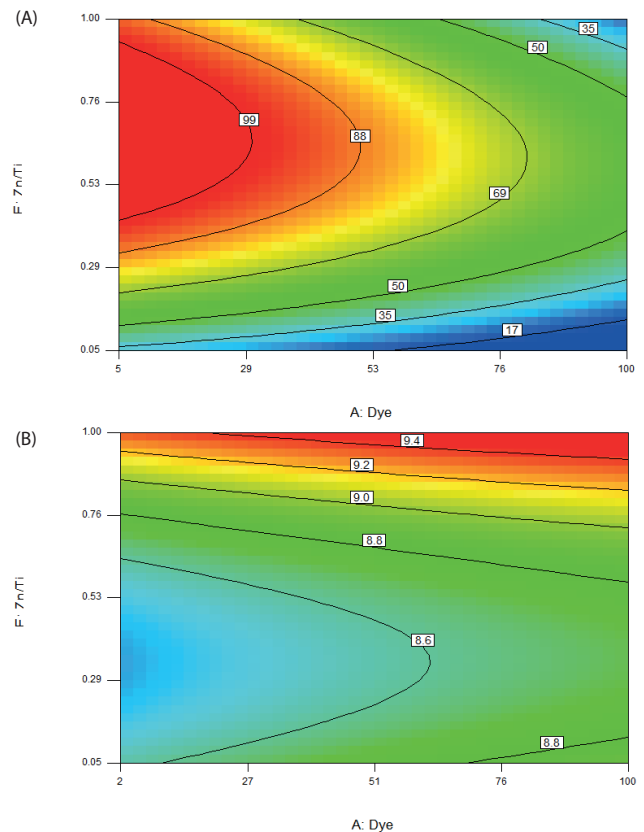


Fig. 6. Effect of initial dye concentration and Zn:Ti: (A) DR and (B) pH-out ($t = 5$ h, $[Zn-TiO_2] = 40$ g/m², pH = 7).

particles decrease the light absorption on the surface of different locations. Jiang et al.'s [28] study reports the same.

3.3.4. Effect of photocatalyst type

The effect of photocatalyst type in constant conditions lasting 6 h, including natural pH dye (6.9), dye concentration of 50 mg/L, and photocatalyst concentration of 44.6 g/m² with a Zn:Ti ratio equal to 0.53, was investigated and it has been shown in Fig. 16. Removal of dye by glue, TiO₂ only, and doped TiO₂ reaches 6%, 25%, and 88%, respectively, which indicates the need for a catalyst and confirms the increase in TiO₂ photocatalytic activity by dopant.

3.4. Optimize the removal conditions

After analyzing the factors affecting the responses, optimization was performed by a central composite design optimization method. The range is determined when color removal above is 90%; dye efficiency degradation is above 67%; and output pH is 9. Fig. 17 shows this range for the natural pH of dye (6.9) at a 6-h duration and 44.6 g/m² photocatalyst.

After finding the optimal conditions to confirm the model, the test was repeated three times under optimum conditions, and average of results are reported in Fig. 18. This condition contained a pH of 6.9, photocatalyst concentration of 44/6 g/m², dye concentration of 53 ppm, and Zn/Ti equal

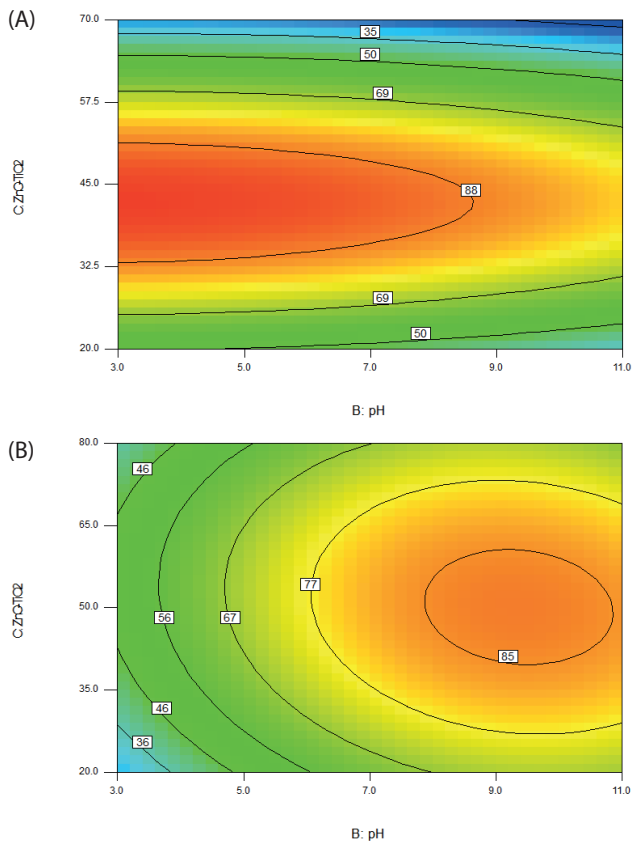


Fig. 7. Effect of amount of photocatalysis and pH: (A) DR and (B) CODR ($t = 5$ h, Zn:Ti = 0.5%, [Dye] = 40 mg/L).

to 53%. According to the diagram, dye removal efficiency was 87% at 6 h and 90% at 8 h, with a 65% organic load removal.

COD removal efficiency increased at first and then decreased and increased again because the dye had three benzene rings. While the matter was not still perfectly under photocatalytic reaction, it was still hard gradable with low chemical oxygen consumption. After the rings broke, the oxygen consumption required for the reaction to decompose increased. By the end, the simple materials that require less oxygen were formed.

During Fig. 18(C), an output pH of 8.5 was observed at 6 h. With increased time, the effect of doped metal on dye removal increases because there is ample opportunity for response. Additionally, more hydroxyl and oxygen radical are produced because the dye investigated has been anion and buffered, and the output pH increases.

The results, including color removal of 88%, organic load removal of 65%, and output pH of 8.5, compare with the predicted responses at the optimal point, as seen in Table 3. The table also shows that the range of average results occur 95% of the time. Therefore, it can be said that the model of color removal and output pH are acceptable with a 95% confidence level.

3.5. Result of liquid chromatography–mass spectrometry

To determine the products under optimal conditions, a liquid chromatography–mass spectrometry (LC–mass) test lasting

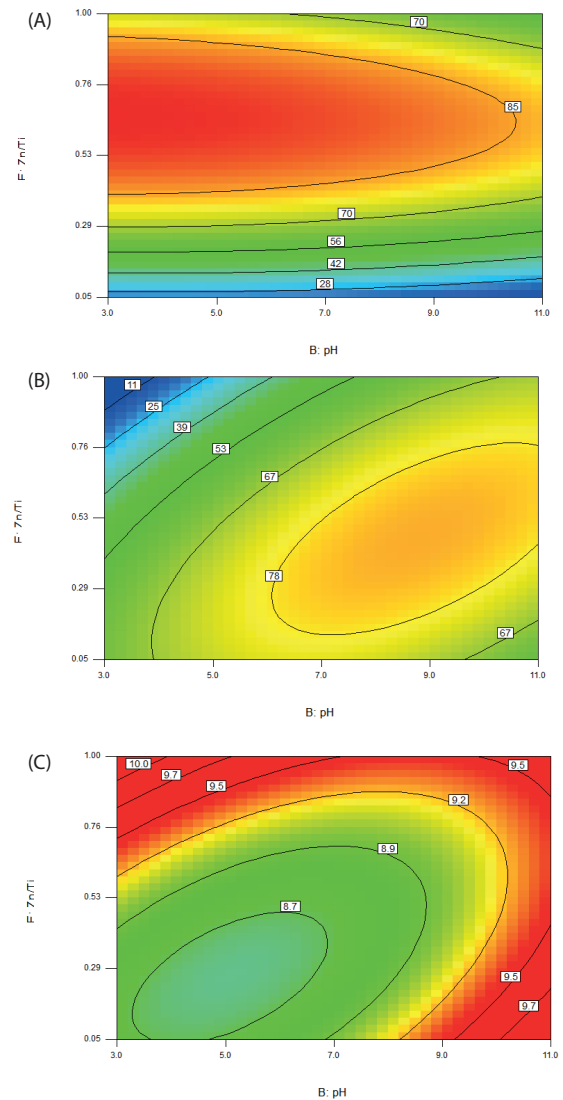


Fig. 8. Effect of pH and Zn:Ti: (A) DR, (B) CODR, and (C) pH-out ($t = 5$ h, [Zn-TiO₂] = 40 g/m², [Dye] = 40 mg/L).

8 h was carried out in optimal conditions to confirm the model. Standardization of the prototype is presented in Figs. 19(A) and (B). In Fig. 19(B), the rate of mass obtained at the peak point of 59 min is equal to 941.1. This number is related to Direct Blue dye in that four sodium ions are lost and four protons (H⁺) are taken. The results of the LC–mass test are shown in Figs. 20(A) and (B). As seen in Fig. 20(A), there is no peak at 5.9 min. Next, the mass at this point was determined, and the result is reported in Fig. 20(B). Peak values in these samples indicate poor residual compounds in a solution. It is safe to say that there are no circular complex combinations in the environments and all ingredients are simple in structure. Over time, the complexity of the reaction products reduces, representing the process' ability to reduce effluent toxicity. Intermediate product types during the reaction depend on process steering conditions, such as the initial concentration and reaction time.

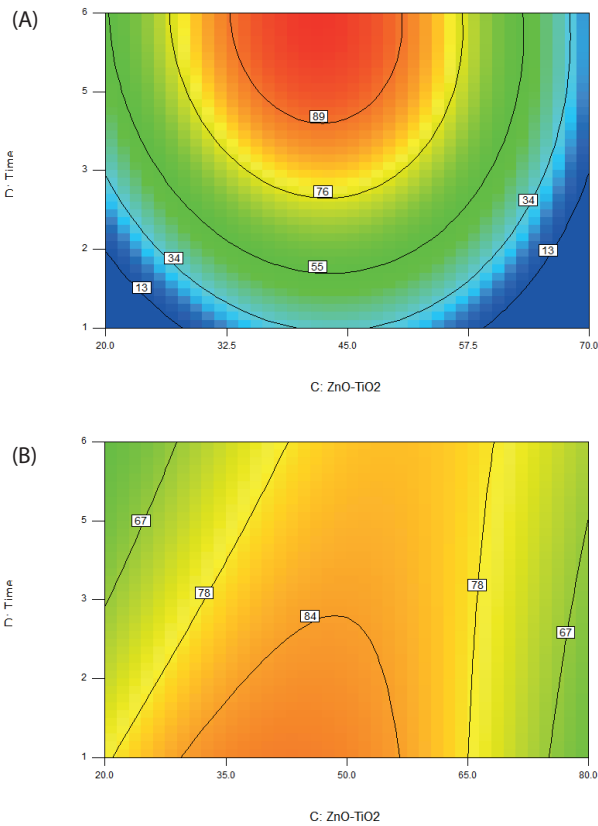


Fig. 9. Effect of time and amount of photocatalysis: (A) DR and (B) COD removed (0.5% Zn-TiO₂, pH = 7, [Dye] = 40 mg/L).

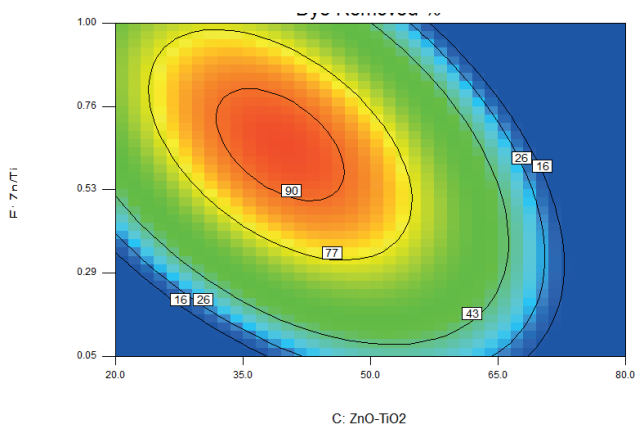


Fig. 10. Effect of Zn:Ti and amount of photocatalysis ($t = 5$ h, pH = 7, [Dye] = 40 mg/L).

4. Conclusion

The XRD test result shows that the phase transition of the anatase to rutile in doped samples is inhibited; meaning the photocatalytic activity of the doped samples is not more than that of the other samples. The removal of reactions in a continuous solar photo-reactor and in Zn-TiO₂ as a photocatalyst was accomplished to analyze the factors affecting efficiency. The test's design was a central composite design, and to ensure the results, the design and experimental testing were

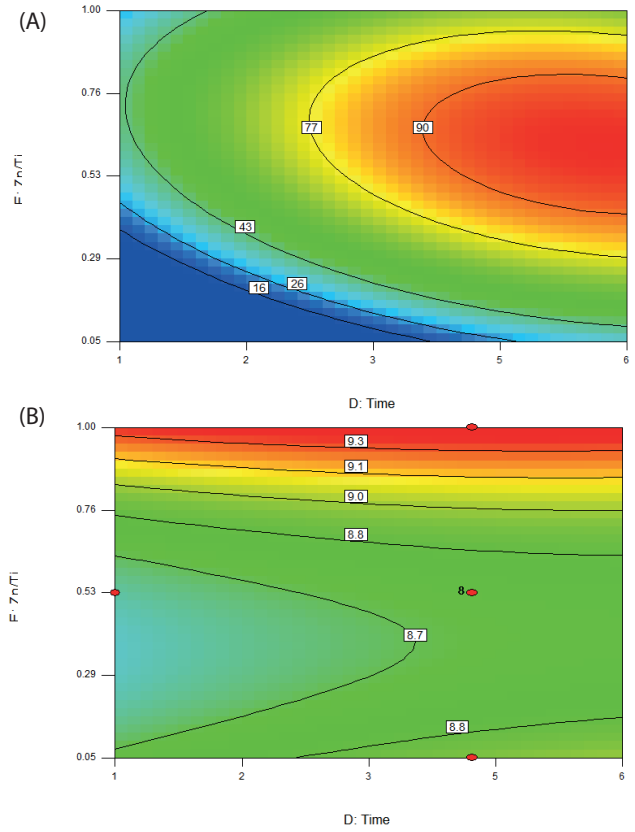


Fig. 11. Effect of Zn:Ti and time: (A) DR and (B) pH-out ([Zn-TiO₂] = 40 g/m², pH = 7, [Dye] = 40 mg/L).

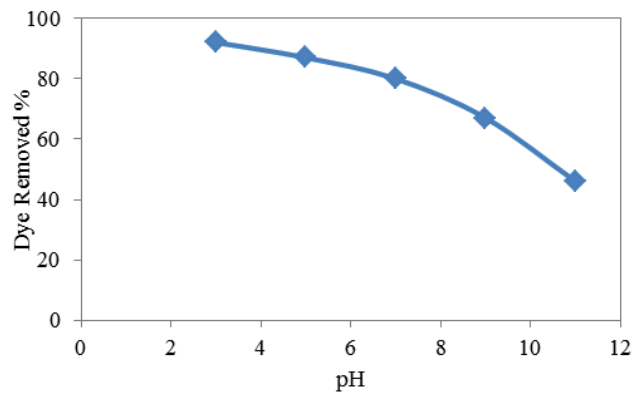


Fig. 12. Effect of pH on dye removal ($t = 5$ h, [0.5% Zn-TiO₂] = 40 g/m², [Dye] = 50 mg/L).

performed once. The effects of parameters such as photocatalyst amount, initial pH, pollutant concentration, and zinc to titanium ratio in the dopant construction phase were studied. It was observed that in optimal conditions, including a photocatalyst of 4.6 g/m², a natural dye pH of 6.9, a dye concentration of 50 mg/L, a duration of 6 h in the sun, and a Zn:Ti ratio equal to 0.53%, dye removal can reach about 90%. Finally, a LC-mass was completed to determine the toxicity of the products produced. The results of these tests indicate destruction of all benzene rings and production of harmless secondary products.

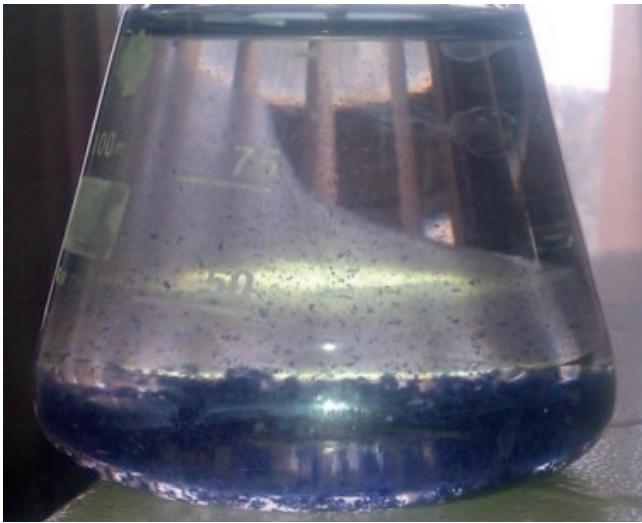


Fig. 13. Clotting dye in acidic conditions.

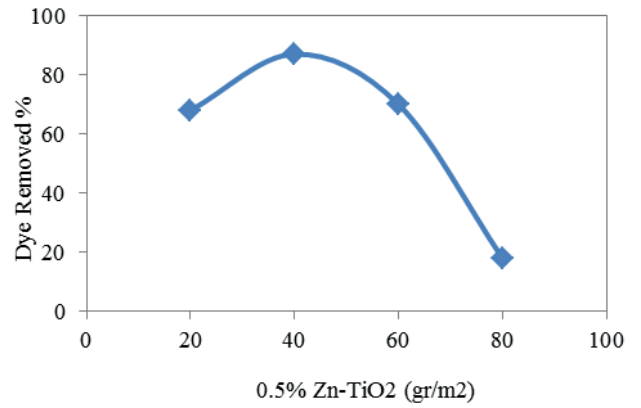


Fig. 15. Effect of photocatalyst concentration on dye removal ($t = 5$ h, [Dye] = 50 g/m², Zn:Ti = 0.5%, pH = 7).

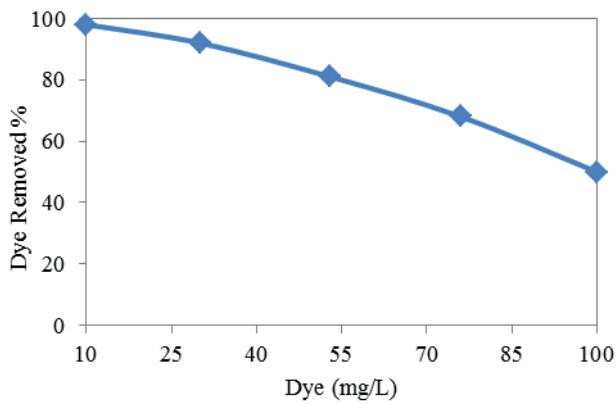


Fig. 14. Effect of dye concentration on dye removal ($t = 5$ h, [0.5% Zn-TiO₂] = 40 g/m², pH = 7).

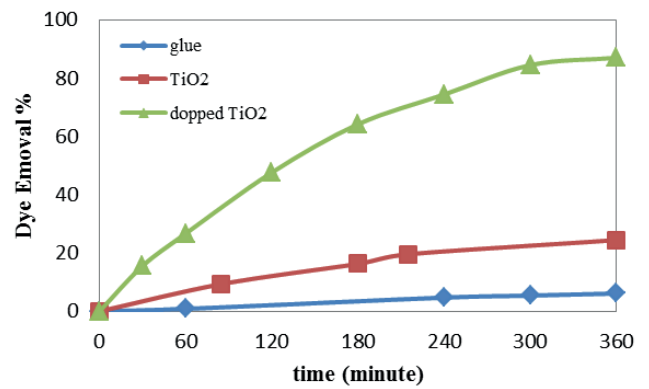


Fig. 16. Effect of photocatalyst type on dye removal ($t = 5$ h, [Dye] = 50 mg/L, [0.5% Zn-TiO₂] = 40 g/m², pH = 7).

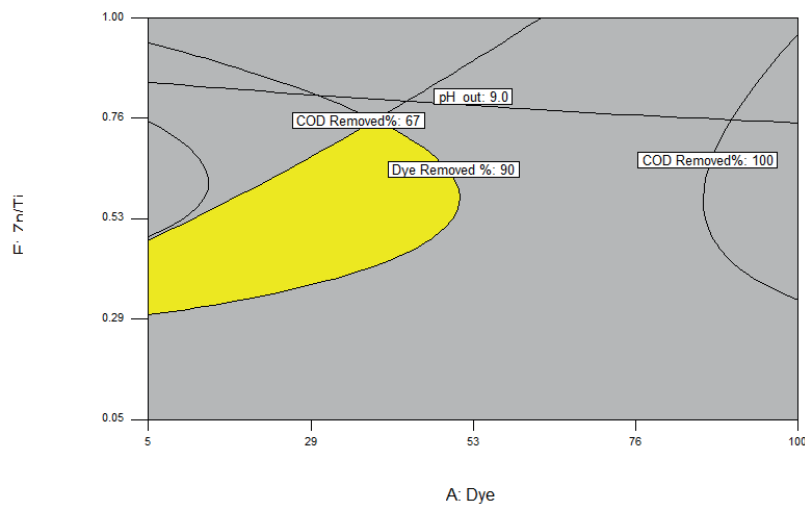


Fig. 17. Range of optimum conditions ($t = 5$ h, [Zn-TiO₂] = 44.6 g/m², pH = 6.9).

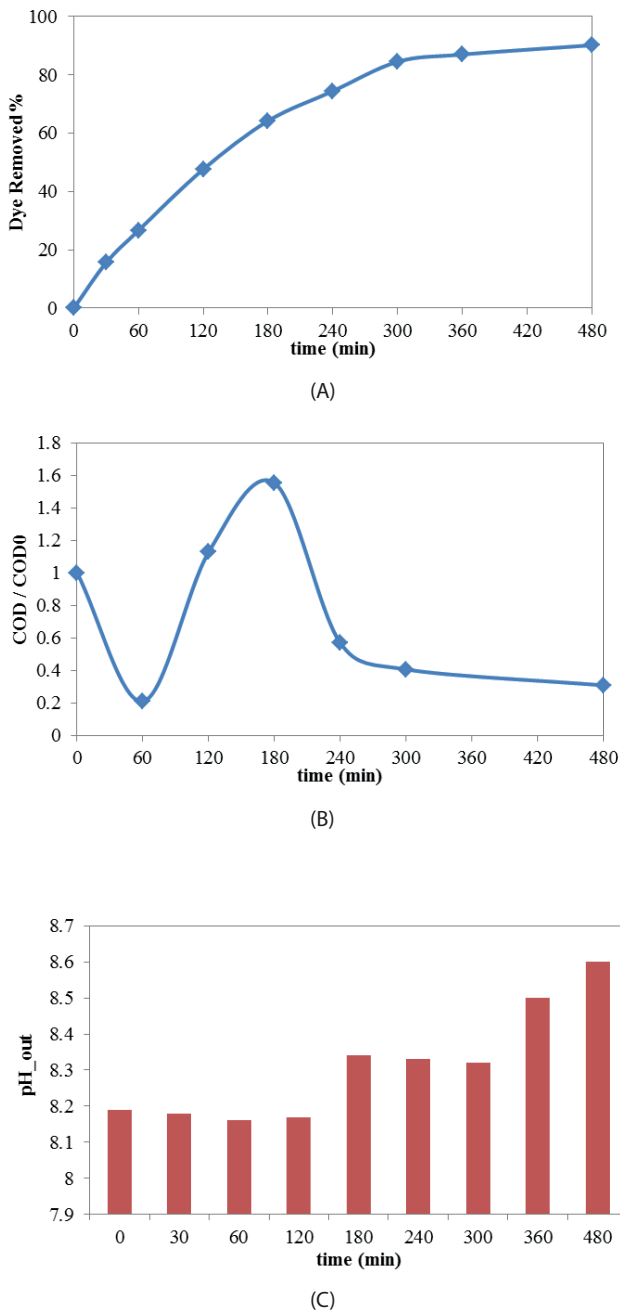
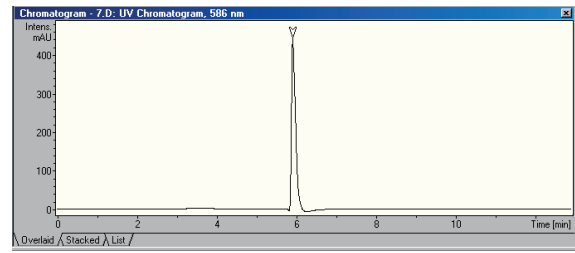


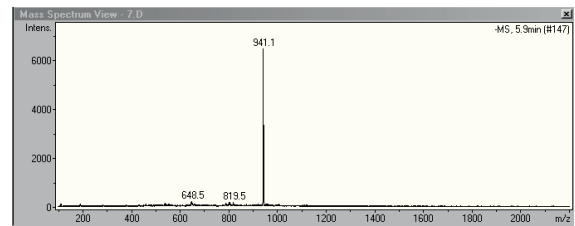
Fig. 18. The results in optimum condition: (A) DR, (B) CODR, and (C) pH-out (pH = 6.9, [0.53% Zn-TiO₂] = 44.6 g/m², [Dye] = 53 mg/L).

Table 3
View optimal point

Result	Prediction	95% Confidence interval (CI) low	95% Confidence interval (CI) high	95% Prediction interval (PI) low	95% Prediction interval (PI) high
Dye removed %	90	83.6	96.6	70	110
COD removed %	84	76.2	92.2	56.4	112
pH-out	8.7	8.5	8.9	8.1	9.2

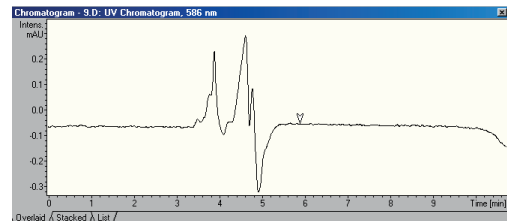


(A)

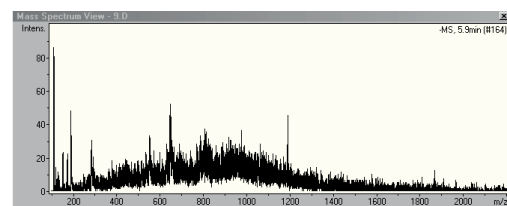


(B)

Fig. 19. Standardization of the pre-sample.



(A)



(B)

Fig. 20. Mass of eighth-hour sample.

Acknowledgment

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