ZnFe-layered double hydroxide nanostructures prepared from natural iron ore precursors with the hydrothermal method and the photocatalytic degradation of Rhodamine B

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\textbf{ABSTRACT}

This paper investigates the use of low-cost natural iron ore in the preparation of binary metal layered double hydroxide (ZnFe-LDH) nanostructures via a hydrothermal process. In particular, the photocatalytic degradation of Rhodamine B under UV light irradiation is utilized. Natural iron ore is used to prepare the iron(III) precursors (Fe\textsuperscript{3+}) necessary for building the ZnFe-LDH nanostructures. Preparation is done at 80°C–100°C for 36 h and does not involve a calcination process. The physicochemical characterizations of the prepared samples are carried out via X-ray diffraction, Fourier-transform infrared spectroscopy, energy-dispersive X-ray spectroscopy, scanning electron microscopy, UV-Vis diffuse reflectance spectroscopy, and photoluminescence. After confirming the existence of the ZnFe-LDH nanostructures via their functional groups and elemental compositions, it is noted that they have high crystallinity and small crystal sizes. The morphologies of the ZnFe-LDH nanostructures are irregular, with small particles agglomerated on their surfaces. While good optical properties are enhanced as the preparation temperature is increased, the results show that crystallinity, crystal size, particle uniformity, and narrow bandgap energy play important roles in the subsequent photocatalytic reactions. The optimum conditions for the photodegradation of 10 ppm Rhodamine B solution are reached with 1.00 g/L of ZnFe-LDH nanostructures prepared using the hydrothermal method at 100°C for 36 h. At maximum efficiency, the photodegradation process can be completed at a rate constant of 0.2060 min\textsuperscript{-1} in 20 min and an \( E_{\text{EO}} \) of 44.74 kWh/m\textsuperscript{3} with the addition of a 0.9% v/v H\textsubscript{2}O\textsubscript{2} electron acceptor. Photocatalyst efficiency decreases by just 12.5% across five test cycles.

\textit{Keywords:} ZnFe-LDH; Iron ore; Hydrothermal; Photocatalytic; Energy consumption
1. Introduction

In recent decades, industrial advancement and increasing urbanization have resulted in various kinds of environmental pollution around the world, especially water pollution [1]. Organic pollutants are known as the main contaminant type of wastewater. Dyes are the most important pollutants that can be produced in the textile, paint, paper, and plastic industries. A significant source of water contamination comes from synthetic dyes in wastewater, such as Rhodamine B (RhB) is a basic dye belonging to a group of heterocyclic organic compounds with a wide range of industrial applications. Human exposure can result in burning eyes, damage to the respiratory system, vomiting, cyanosis, rapid heartbeat, tissue necrosis, nausea, and diarrhea [2]. In addition, it causes serious environmental problems due to its high toxicity and accumulation in the environment. Therefore, water purification is gaining widespread scientific interest [3,4]. There are several methods used to remove impurities, such as adsorption, oxidative processes, electrochemical degradation, and photocatalytic reactions. Among many other techniques, the photocatalytic process has attracted a lot of attention. The advantages of photocatalytic treatment include its flexibility to respond to environmental variables, the use of oxygen that has been dissolved in aqueous media for operations, and the complete mineralization of organic contaminants into more benign by-products [5].

In this type of photocatalytic reaction, a photocatalyst is irradiated, and electron (e–)/hole (h+) pairs are photogenerated on its surface. These pairs form highly reactive species, such as hydroxyl radicals (OH•), superoxide anion radicals (O2−), and H2O2, which induce redox reactions to degrade organic pollutants [6]. Recently, photocatalysts of various semiconductors have been developed for enhanced photocatalysis. Therefore, researchers are focused on the development of new and efficient semiconductors. Various types of photocatalysts are constantly being researched and developed. But there is still a fundamental need to deliver an efficient and useful catalyst through a cheap, easy, and environmentally-friendly route. Most importantly, a semiconductor must be cost-effective to meet future energy and environmental demands [7,8].

One of the new materials for environmental treatment applications is layered double hydroxides (LDHs) because LDHs are good adsorbents and also have good photocatalytic properties such as a high specific surface area, easy preparation, low cost, and high stability [9]. LDHs are well known as a class of two-dimensional anionic clays composed of positively charged metal hydroxide sheets and charge compensating interlayer anions. These are generally expressed by the classical formula [Mn+x/z+(Mn+x/z+(OH)x)‧nH2O, where Mn+x/z+ represents a divalent cation such as Mg2+, Zn2+, Ni2+, Ca2+, etc., and Mm+x/z+ represents a trivalent cation, for example, Al3+, Fe3+, Cr3+, Mn3+, etc. A+x/z stands for the interlayer anions of valence n (CO3–, SO4–, NO3–, Cl–, or PO4–, etc.). x is the molar ratio of Mm+x/z+/Mn+x/z+, and n is the number of water molecules present in the layered structure [10–12]. LDHs are the popular inorganic hosts for the formation of nanocomposite or nanolayered composite materials, and they have received much attention in the past decades due to their vast applicability in areas such as nanotechnology, biotechnology, adsorption, catalysis, and photocatalysis [13].

Natural iron ore predominantly contains hematite (α-Fe2O3), a rich and inexpensive mineral that can be prepared as a photocatalytic material [14]. Fe2O3 is the cheapest, most stable, and most environmentally-friendly semiconductor material, but it has limitations for its photocatalytic activity, such as a high electron–hole recombination rate, weak conductivity and low diffusion length of the hole, and low conductivity [15]. The photocatalytic properties of Fe2O3 can also be enhanced with composites with other metals such as zinc [16]. For efficiency and cost-effectiveness, ZnFe-LDH can be synthesized at a lower heat consumption than oxide for iron/zinc composites [17]. As mentioned above, ZnFe-LDH is attractive as a photocatalyst material due to its low cationic toxicity and compatibility of Zn2+ and Fe2+ [18]. An example of the application of ZnFe-LDH as a photocatalyst for environmental remediation, such as the degradation of aqueous methyl violet (MV) and malachite green (MG) [19], the removal of Microcystis aeruginosa (algae) [20], and the oxidation of ofloxacin (antibiotics) [21], etc. A binary metal ZnFe-LDH has been successfully prepared by numerous methods and precursors [22–24]. These precursors can be good candidates for the synthesis of ZnFe-LDH nanostructures. The preparation and benefits of nanocomposites or nanostructures have been reported several times for their outstanding applications and can act as a suitable photocatalytic agent. They can potentially exhibit significant advantages in environmental applications, such as wastewater purification using nanomaterials and/or nanocomposites [25,26].

Herein, we report on the novel natural iron ore precursors for facile simple one-step hydrothermal preparation of ZnFe-LDH nanostructures. The goal of this research was to increase the value of natural low-cost material in a simple and effective process that could be used as a guideline for large-scale production. This research involved preparing ZnFe-LDH nanostructures from natural iron ore precursors with the chemical precipitation-hydrothermal method and testing for the photocatalytic degradation of RhB from an aqueous solution. Their structural and optical properties have been studied by X-ray diffraction (XRD), Fourier-transform infrared spectroscopy (FT-IR), energy-dispersive X-ray spectroscopy (EDXS), scanning electron microscopy (SEM), UV-Vis diffuse reflectance spectroscopy (UV-DRS), and photoluminescence spectroscopy (PL) techniques. The effects of the photocatalyst preparation temperature, the electron acceptor concentration for the reaction efficiency, the energy consumption of the photodegradation process, and the photocatalyst reusability were studied.

2. Materials and methods

2.1. Materials

The natural iron ore in this experiment was obtained from natural resources in the Namphee sub-district, Thongsaenkhan District, Uttaradit Province, Thailand. The other chemicals were analytical grade reagents, including hydrochloric acid (HCl, 35.4%, Loba Chemie, India), hydrogen peroxide (H2O2, 30% v/v, Merck, Germany), zinc acetate dihydrate (Zn(CH3COO)2·2H2O, 99.5%, Loba Chemie,
India), sodium hydroxide (NaOH, 98%, Loba Chemie, India), Rhodamine B (C₂₀H₂₈ClN₂O₃, 95%, Loba Chemie, India), isopropanol (C₃H₈O, 99.5%, Loba Chemie, India), ethylene diamine tetra-acetic acid (C₁₀H₁₆N₂O₈, 99.4%, Loba Chemie, India), and L-ascorbic acid (C₆H₈O₆, 99.5%, Loba Chemie, India). All of the solutions were prepared with deionized (DI) water.

2.2. Preparation of natural iron ore

Natural iron ore is characterized by small granules of magnetite (Fe₃O₄) mixed with reddish-brown hematite powder. In this research, natural iron ore was sifted through a 200-mesh sieve to collect only the powdered hematite for use in the subsequent preparation of the Fe²⁺ solution. 20 g of sifted iron ore powder were mixed with 100 mL of 6 M HCl and stirred for 30 min, followed by hydrothermal treatment at 100°C for 10 h under autogenous pressure. After cooling, filtration was used to separate the suspension and retain only parts of the solution. After that, 10 mL of concentrated H₂O₂ was added into a solution for the oxidation of Fe²⁺ to Fe³⁺ ions. The concentration of Fe³⁺ ions was calculated using the titration method.

2.3. Preparation and characterization of ZnFe-LDH photocatalysts

The ZnFe-LDH nanostructures were prepared with the 0.05 M zinc acetate and 0.05 M Fe³⁺ solutions described in the previous section. These two solutions were slowly mixed together. Then 6 M of NaOH solution was added dropwise to the mixture under continuous and vigorous stirring until precipitation was complete. Once the solution was homogeneous, the pH of the reaction was alkaline (pH~10.0) and precipitates were separated from the reaction medium via filtration and then thoroughly washed with water. The precipitates were separated from the reaction medium via filtration and then thoroughly washed with water.

The final samples were characterized via X-ray powder diffraction (Bruker/D8 Phaser, Germany) with CuKα radiation (λ = 1.5418 Å) while scanning from 10° to 80° at a rate of 0.02°/s. Fourier transform infrared spectroscopy (Perkin Elmer Spectrum RX I, UK), energy-dispersive X-ray spectroscopy (Oxford Instruments, Ultim Max 40, UK), scanning electron microscopy (TESCAN VEGA3, Czechia), UV-Vis diffuse reflectance spectroscopy (Shimadzu UV-3101PC, Japan), and photoluminescence spectroscopy (Avantes AvaSpec-2048TEC, Netherlands) were also utilized.

2.4. Photocatalytic studies

The photocatalytic set-up consisted of 50 mL test tube glass reactors, an initial RhB concentration of 10 ppm, 0.050 g of photocatalyst, and two six-watt UV-C tube lamps (Sylvania, G6W, Japan), as shown in Fig. 1. The UV lamps were utilized due to their high effectiveness and low energy consumption. In each test, chemical adsorption-desorption equilibrium was achieved through stirring continuously for 30 min, then letting the system sit for 30 min in a black-walled box without external illumination. Next, 3 mL samples were collected for evaluation at approximately 0, 5, 10, 15, 20, 30, and 40 min. Finally, the samples were filtered through a 0.22 μm syringe filter (MS® MCE Syringe filter, Membrane Solution, USA). These tests were then used to create a guideline for developing future treatment systems.

The absorbance of the RhB solution was measured with a UV-Vis spectrophotometer (T90+, PG Instruments, UK). More details can be found in our previous report [27,28]. The degradation efficiency (DE (%)) was calculated using the following equation [21]:

\[
\text{DE} (%) = \left[ 1 - \left( \frac{A_t}{A_0} \right) \right] \times 100
\]

(1)

where \(A_0\) and \(A_t\) are the initial and final absorbances of the RhB solution at 0 min and \(t\) min, respectively.

2.4.1. Active species trapping test

To find out the reactive species in the synthesized ZnFe-LDH photocatalysts. The photocatalytic degradation of RhB over the ZnFe-LDH photocatalysts under UV light irradiation enabled the active species trapping tests. The experiment was monitored after the addition of 4 mM solutions of isopropanol (IPA), ethylenediaminetetraacetic acid (EDTA), and L-ascorbic acid (AA) as the scavengers for OH•, h⁺, and O₂•⁻, respectively.

2.4.2. Effect of the electron acceptor concentration

With the effect of the electron acceptor concentration, H₂O₂ was used as an electron acceptor to evaluate the...
photocatalytic activity of the prepared samples. Different concentrations of H$_2$O$_2$, approximately 0.0% v/v, 0.3% v/v, 0.6% v/v, and 0.9% v/v, were studied with the most efficient photocatalyst.

2.4.3. Reusability test

The reusability of each photocatalyst was investigated by repeatedly degrading the fresh RhB solution (10 ppm) across five test cycles. After each typical photocatalytic experiment, the photocatalyst was separated out via centrifugation and then washed with 10 mL of ethanol for 1 h under magnetic agitation. Then the recycled photocatalyst was reused under the same conditions.

3. Results and discussion

3.1. Characterization of ZnFe-LDH photocatalysts

3.1.1. XRD analysis

The X-ray diffraction patterns of the samples prepared with the hydrothermal method are shown in Fig. 2. As illustrated in Fig. 2a, all of the XRD patterns show major peaks at $2\theta = 24.1^\circ$, $33.2^\circ$, $35.7^\circ$, $40.8^\circ$, $49.5^\circ$, $54.1^\circ$, $62.5^\circ$, $64.1^\circ$ and $72.6^\circ$, which correspond to the (012), (104), (110), (113), (024), (116), (214), (300), and (119) peaks of rhombohedral $\alpha$-Fe$_2$O$_3$, in accordance with the Joint Committee on Powder Diffraction Standards (JCPDS) file No. 00-033-0664 [29]. Fig. 2b shows the main characteristic peaks at the 20 values of $31.9^\circ$, $34.6^\circ$, $36.4^\circ$, $47.7^\circ$, $56.8^\circ$, $63.0^\circ$, $66.5^\circ$, $68.1^\circ$, $69.2^\circ$, and $72.8^\circ$, corresponding to the (100), (002), (101), (102), (103), (200), (112), (201), and (204) diffraction planes of hexagonal ZnO, respectively (JCPDS file No. 01-079-0205). All of the reflections of the sample prepared at $80^\circ$C (Fig. 2c) corresponded to crystalline ZnO. The peaks representing the other phases were not yet apparent in this sample. For the samples prepared at $90^\circ$C (Fig. 2d), the reflection peaks around $11^\circ$, $24^\circ$, $35^\circ$, $39^\circ$, $47^\circ$, $59^\circ$, and $61^\circ$ corresponded to the reflections of planes hkl (003), (006), (009), (015), (012), (110), and (113) of Zn–Fe layered double hydroxide (ZnFe-LDH), which were the typical characteristic peaks of hydrotalcite [20]. At higher preparation temperatures (Fig. 2e), the signal intensity of the characteristic peaks of the ZnFe-LDH also increased. As well, the main peaks of the prepared ZnFe-LDH occur at the 20 angles of $11^\circ$ and $24^\circ$ (Fig. 2d and 1e). The peak found at $32^\circ$ can be attributed to the ZnO phase, which forms when there are a large number of Zn(OH)$_2$ crystallization centers in the brucite-like layer [20]. The prepared samples at $90^\circ$C–$100^\circ$C were observed to have a successful formation for the ZnFe-LDH crystal phase. The signal intensity of the (003) and (006) peaks (LDH peaks) was increased with the increase of the preparation temperature. A higher crystallinity of the (003) and (006) peaks was observed, which indicated the presence of a well-ordered layered structure [30]. The average crystallite size was calculated by employing Scherrer’s formula [31]:

$$D = \frac{0.9\lambda}{\beta \cos \theta}$$

where $D$, $\lambda$, $\beta$, and $\theta$ represent the crystallite size, the wavelength of the X-ray used, the full width at half maximum (FWHM, in radians), and the Bragg diffraction angle, respectively. The average crystallite sizes of the prepared ZnFe-LDH obtained at different preparation temperatures were 20.54 and 17.15 nm for ZnFe-LDH prepared at $90^\circ$C and $100^\circ$C, respectively. The smaller crystallite size of the particles implies a larger photocatalyst surface area. The ZnFe-LDH prepared at $80^\circ$C did not show major LDH peaks, so the crystal size could not be calculated. The small crystal sizes in the ZnFe-LDH enable the rapid transfer of electrons from the valence band to the conduction band. They also inhibit electron–hole recombination, leading to enhanced photocatalytic activity [32]. Crystallite size is one of the important factors that have a vital role in photocatalytic activity [19].

3.1.2. FT-IR analysis

The FT-IR spectra of samples prepared at $80^\circ$C–$100^\circ$C for 36 h in the range of 400–4,000 cm$^{-1}$ at room temperature are shown in Fig. 3. The absorption bands below 1,000 cm$^{-1}$ were the M–OH and M–O vibration modes of ZnFe-LDH [33,34]. A broad strong absorption band centred at 3,420 cm$^{-1}$ was attributed to the stretching vibrations of the surface and interlayer water molecules and hydroxyl groups in the LDH layers, which were found at lower frequencies in the LDH compared with the O–H stretching vibration in free water at 3,600 cm$^{-1}$.

Fig. 2. X-ray diffraction patterns of (a) Fe$_2$O$_3$, (b) ZnO, and ZnFe-LDH samples prepared with the hydrothermal method at (c) $80^\circ$C, (d) $90^\circ$C, and (e) $100^\circ$C.
The bands at 1,634 cm$^{-1}$ could be attributed to the bending vibration of the water molecules adsorbed between them [36]. Other peaks, such as the peak band at approximately 2,850–2,950 cm$^{-1}$ of the C–H stretch vibration [37], were likely acetate ions from the precursors. The Cl$^-$ vibration at 830 cm$^{-1}$ might have been from the precursors of the natural iron ore preparation process [19,20]. The FT-IR patterns of all three samples prepared at 80°C–100°C were very similar and the peak intensity was increased by the preparation temperature. Based on these similar peaks, it was more probable that the one prepared at 80°C was ZnFe-LDH.

3.1.3. EDXS analysis

EDXS analysis was also employed to ascertain the chemical composition of the selected area on the surface of the raw materials and ZnFe-LDH nanostructures, as shown in Fig. 4. For the inset of the EDXS spectra, we report the chemical composition detected with the relative percentage in weight (wt.%). The main element composition of the natural iron ore consisted of Si, Fe, and Al (Fig. 4a). Other elements, such as Au and C, were generated by the analytical equipment and process. After the iron ore preparation process for Si removal, the main remaining element was Fe and a small amount of Al was left (Fig. 4b). When the ZnFe-LDH was prepared at various hydrothermal reaction temperatures, lattice oxygen and Cl anions were correctly intercalated between the metallic layers (Fig. 4d–f). The percentage of Fe by weight increased as the reaction temperature increased in the range 80°C to 100°C, and the M$^{2+}$/M$^{3+}$ molar ratio decreased from 3.15 to 2.12 in that same range. According to the literature, if the molar ratio is too high, then a low substitution rate for M$^{3+}$ in the M$^{2+}$(OH)$_2$ layers induces the formation of LDH with strongly reducing porosity [38]. Hence, a low M$^{2+}$/M$^{3+}$ ratio is preferred for LDH synthesis. In previous works, the Zn/Fe ratio has been shown to affect the structural properties and performance of the LDH.

Fig. 3. Fourier-transform infrared spectra of ZnFe-LDH nanostructures prepared with the hydrothermal method at (a) 80°C, (b) 90°C, and (c) 100°C.

Fig. 4. Energy-dispersive X-ray spectra of (a) iron ore, (b) Fe$_2$O$_3$, (c) ZnO, and ZnFe-LDH nanostructures prepared with the hydrothermal method at (d) 80°C, (e) 90°C, and (f) 100°C.
3.1. SEM analysis

The particle size and surface morphology of the samples were evaluated using SEM microscopy. Fig. 5 shows the SEM micrograph of the prepared samples with 10,000× magnification. The morphology of the natural iron ore (Fig. 5a) was a rectangular lump with small particles scattered on the surface and particle sizes of approximately 30 µm × 30 µm. The morphology of the pure Fe₂O₃, as shown in Fig. 5b, was a uniformly distributed spherical particle with a diameter of approximately 0.5–1.5 µm. The pure ZnO Fig. 5c had uniform small rod particles with sizes of approximately 1.0–2.0 µm. The morphology of the ZnFe-LDH nanostructures prepared from natural iron ore (Fig. 5d-f) showed that all three samples were irregular in shape, with tiny particles clumping on the surface [1,21]. At higher preparation temperatures, the surface-bound particle size was also larger. These results confirmed the successful formation of ZnFe-LDH particles.

3.1.5. Optical analyses

The efficiency of a photocatalyst involves light absorption and light-induced electrons and holes, which are the key factors controlling a photocatalytic reaction. The electronic excitation of a photocatalyst after the absorption of energy from the light source can be calculated to determine the band gap energy (Eg).

\[
(ahv)^{1/n} = K(hv - E_g)
\]

(4)

where index n determines the type of transition, that is, it takes the values 1/2, 2, 3/2, or 3 for directly allowed, indirectly allowed, forbidden direct, and forbidden indirect transitions, respectively. Hence, for direct transitions of ZnFe-LDH nanoparticles, n is 1/2 [40]. Fig. 6d-f exhibits the bandgap energies of the ZnFe-LDH nanostructures, which were found to be 2.08, 2.00, and 1.95 eV for ZnFe-LDH80, ZnFe-LDH90, and ZnFe-LDH100, respectively. These bandgaps are narrower than those for ZnFe-LDH prepared via chemical co-precipitation without a hydrothermal step [21]. The bandgap energies were decreased when the preparation temperature was increased. This implied that the photo-absorption of ZnFe-LDH100 was strongest, which resulted in the highest photocatalytic activity [19].

Fig. 7 shows the PL spectra of the ZnFe-LDH nanostructures, which were analysed to confirm the optical properties of the samples. The PL spectra of all the ZnFe-LDH nanostructures activated at 590 nm showed sharp radiation peaks at 665 nm, with the peak intensity increasing with the increasing preparation temperature. This revealed more separation capacity for the e/h⁺ pairs and the formation of OH⁻ [21]. However, the hydroxide groups present on the surface of the ZnFe-LDH nanostructures contributed to the reduction of the e/h⁺ recombination, resulting in better photocatalytic activity.

3.2. Photocatalytic activity

The photocatalytic behaviour and the pseudo-first-order kinetic plot of the prepared ZnFe-LDH nanostructures were evaluated using the degradation of the RhB in an aqueous solution under UV light irradiation, as shown in Fig. 8. The photocatalytic efficiency and the kinetic rate constant values of the samples are shown in Table 1. The highest efficiency for the photocatalytic degradation of RhB was for the ZnFe-LDH100 catalyst with a degradation efficiency of 68.53%, and a kinetic rate constant of 0.0348 min⁻¹ in 40 min. Based on this result, the photocatalytic efficiency was increased by increasing the preparation temperature. This was because increasing the preparation temperature resulted in better photocatalytic properties such as higher crystallinity, reduced crystal size, and reduced bandgap energy. The preparation of LDH using the hydrothermal method could be easier for tuning structures and morphologies [7]. Radical chemistry plays an important role in photocatalysis. The quantity and the chemical nature of the radicals generated in a particular photocatalytic reaction are highly dependent on the surface chemistry and electronic properties of the catalyst [42].

To confirm the main active species and the potential reaction mechanism for superior photodegradability of the ZnFe-LDH photocatalyst, a trapping experiment was carried out without the electron acceptor added. In this study, IPA, EDTA, and AA were added to the reaction solutions as OH⁻, h⁺, and O₂⁻ scavengers, respectively [43,44]. The results are illustrated in Fig. 9. When the different scavengers are participated in the photocatalytic reaction, the RhB degradation performance is obviously reduced. The calculated RhB dye degradation efficiency of 23.12%, 19.37%, and 62.85% correspond to the scavengers IPA, EDTA, and AA, respectively. EDTA and IPA scavenger observed a major decline in the degradation efficiency, which inferred that h⁺ and OH⁻ play a key role in the photocatalytic process. Based on this experiment, the role of active species in the photocatalytic degradation follows the order as h⁺ > OH⁻ > O₂⁻.

Fig. 10 shows the degradation efficiency of the ZnFe-LDH100 samples for the different concentrations of H₂O₂. For the UV light-induced degradation, percentages in the
range of 68.53%–100.00% with a kinetic rate constant in the range of 0.0348–0.2060 min\(^{-1}\) were observed for \(\text{H}_2\text{O}_2\) concentrations of 0.0%–0.9% (Table 1). The highest efficiency was that for added 0.9% v/v \(\text{H}_2\text{O}_2\) with complete photodegradation in 20 min. The \(\text{H}_2\text{O}_2\) acted as an electron acceptor that prevented \(e^-/h^+\) pairs recombination as it trapped photo-generated electrons and increased the availability of holes in the generation of reactive oxygen species (ROS) \[45\]. However, the addition of \(\text{H}_2\text{O}_2\) to the photocatalytic reaction could enhance the degradation efficiency by reacting with an electron to produce ROS as well. This might have been partly due to the presence of Fe in the ZnFe-LDH sample, which

Fig. 5. Scanning electron micrograph of (a) iron ore (b) Fe\(_2\)O\(_3\) (c) ZnO, (d) ZnFe-LDH80, (e) ZnFe-LDH90, and (f) ZnFe-LDH100.
could react with H$_2$O$_2$ to produce more hydroxyl radicals and exhibit better photocatalytic activity [13].

Energy consumption is one of the various factors that can affect the selection of photocatalytic processes since it affects the cost-effectiveness of the process. The electrical energy per order ($E_{EO}$) was the parameter introduced in this research to evaluate advanced oxidation processes (photocatalytic process with UV). $E_{EO}$ has been proposed as an indicator of process efficiency in terms of cost per unit volume for UV reactors. The following equation was used to calculate the $E_{EO}$ values in kWh/m$^3$ order given the first-order degradation reaction [46,47]:

$$E_{EO} = \frac{38.4 \times P}{V \times K}$$  \hspace{1cm} (5)

where $P$ is the power of the light source (kW) and $V$ is the solution volume (L). As displayed in Table 1, the calculated $E_{EO}$ values ranged from 264.82 to 708.92 kWh/m$^3$ for the case without an electron acceptor. In terms of the energy consumption in the photocatalytic degradation process, it was found that more than two times the energy was to be able to be saved when ZnFe-LDH100 was used as a photocatalyst compared to photolysis. For the case of an electron acceptor being added, the calculated $E_{EO}$ was in the range of 44.74–264.83 kWh/m$^3$, which was found to decrease as the
concentration of the H$_2$O$_2$ electron acceptor in the photocatalytic reaction increased. Based on the experimental results, the energy consumption could be reduced by approximately six times when the H$_2$O$_2$ electron acceptor concentration was increased from 0.0% to 0.9%. The addition of an electron acceptor could reduce the energy consumption of the photocatalytic processes. This was consistent with the previous report of Elbadawy et al. [48]. It was found that the addition of the electron acceptor reduced the energy consumption from 714.29 to 117.6 kWh in the UV/TiO$_2$/ZnO/H$_2$O$_2$ system at pH = 6.5. Energy consumption is a cost-effective indicator for a photocatalytic process that depends on the kinetic rate of the reaction. $E_{EO}$ is considered a key factor in the selection of photocatalytic processes, but there have been no previously reported studies on the energy consumption of ZnFe-LDH photocatalysts.

Fig. 11 shows the five cycles of reusability tested with the photocatalytic degradation of RhB under UV light. A total of five test cycles showed the degradation efficiency for each cycle, as follows. The first two cycles could be completely degraded in 20 and 40 min. The efficiencies of the last three cycles were decreased by 4.17%, 8.33%, and 12.50%. The results revealed that the efficiency exhibited a 12.5% decrease in the total efficiency after five cycles were run, which could have resulted from the photo-corrosion phenomenon [49]. The ZnFe-LDH photocatalyst prepared from low-cost natural iron ore was reasonably stable.

According to the experimental results, a possible mechanism for the photocatalytic degradation of RhB using ZnFe-LDH under UV light irradiation was proposed as shown in Fig. 12.

The primary mechanism for the photocatalytic degradation of the RhB solution is presented below [20,50]. The ZnFe-LDH photocatalysts could be irradiated with light with energy greater than their bandgap, thus producing ZnFe-LDH ($e^{-}_{CB} + h^{+}_{VB}$) pairs in the conduction and valence bands [Eq. (6)].

$$
\text{ZnFe-LDH} + h^{\frac{1}{2}} \rightarrow \text{ZnFe-LDH} \left( e^{-}_{CB} + h^{+}_{VB} \right)
$$

The ZnFe-LDH ($h^{+}_{VB}$) was oxidized with H$_2$O$_2$, yielding OH$\cdot$, as shown in Eqs. (7) and (8), while the ZnFe-LDH ($e^{-}_{CB}$) reacted with O$_2$ to produce O$_2$$\cdot^-$, as illustrated in Eq. (9).

$$
\text{ZnFe-LDH} + \text{H}_2\text{O}_2 \rightarrow \text{ZnFe-LDH} + \text{OH}^\cdot + \text{H}_2\text{O}
$$

$$
\text{ZnFe-LDH} + \text{O}_2 \rightarrow \text{ZnFe-LDH} + \text{O}_2^\cdot\cdot
$$

Table 1

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<th>Samples</th>
<th>Efficiency (%)</th>
<th>$K$ (min$^{-1}$)</th>
<th>$R^2$</th>
<th>$E_{EO}$ (kWh/m$^3$)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Photolysis</td>
<td>36.53</td>
<td>0.0130</td>
<td>0.9818</td>
<td>708.92</td>
</tr>
<tr>
<td>Fe$_2$O$_3$</td>
<td>51.91</td>
<td>0.0210</td>
<td>0.9712</td>
<td>438.85</td>
</tr>
<tr>
<td>ZnO</td>
<td>62.35</td>
<td>0.0238</td>
<td>0.9585</td>
<td>387.22</td>
</tr>
<tr>
<td>ZnFe-LDH80</td>
<td>55.12</td>
<td>0.0292</td>
<td>0.9554</td>
<td>315.61</td>
</tr>
<tr>
<td>ZnFe-LDH90</td>
<td>62.06</td>
<td>0.0300</td>
<td>0.9465</td>
<td>307.20</td>
</tr>
<tr>
<td>ZnFe-LDH100</td>
<td>68.53</td>
<td>0.0348</td>
<td>0.9465</td>
<td>264.82</td>
</tr>
<tr>
<td>ZnFe-LDH100 + 0.3%H$_2$O$_2$</td>
<td>71.00</td>
<td>0.0373</td>
<td>0.9459</td>
<td>247.08</td>
</tr>
<tr>
<td>ZnFe-LDH100 + 0.6%H$_2$O$_2$</td>
<td>100.00</td>
<td>0.0940</td>
<td>0.9961</td>
<td>98.04</td>
</tr>
<tr>
<td>ZnFe-LDH100 + 0.9%H$_2$O$_2$</td>
<td>100.00</td>
<td>0.2060</td>
<td>0.9993</td>
<td>44.74</td>
</tr>
</tbody>
</table>

Fig. 8. The photocatalyst comparisons of the samples prepared with the hydrothermal method, (a) degradation efficiency and (b) pseudo-first-order kinetic plot.

Fig. 9. Active species trapping experiment over ZnFe-LDH100 photocatalyst.
Additionally, the $O_2^*$ could react with $H^+$ and ZnFe-LDH (e$_{cb}$) to produce $OH^*$, as shown in Eqs. (10)–(12). The $H_2O_2$ added to the photocatalytic reaction could react according to Eqs. (12) and (13), reducing the recombination rate and increasing the ROS yield.

$$O_2^* + H^+ \rightarrow HOO^*$$  

$$ZnFe – LDH(e_{cb}) + HOO^* + H^+ \rightarrow H_2O_2$$  

$$ZnFe – LDH(e_{cb}) + H_2O_2 \rightarrow OH^* + OH^-$$  

$$H_2O_2 + h\frac{\lambda}{2} \rightarrow 2OH^*$$

All of the $OH^*$, $O_2^*$, and $H_2O_2$ formed in the photocatalytic reaction resulted in RhB degradation [Eqs. (14) and (15)]. The degradation of RhB involves a de-ethylation
process. It has been observed that each de-ethylated step results in a blue shift of the maximum absorption band [51].

\[ \text{RbB} + \text{OH}^* \rightarrow \text{degradation product} \quad (14) \]

\[ \text{RbB} + \text{O}_2^* \rightarrow \text{degradation product} \quad (15) \]

Based on this work, there are several benefits in using ZnFe-LDH in photocatalysis. First, the preparation process can save energy by using a temperature below that used for the preparation of mixed metal oxides without a calcination process [52]. Second, the photocatalyst is inexpensive because it uses a precursor made from natural iron ore. Finally, it possesses good photocatalytic properties, including high crystallinity, small crystals, and uniform particles, and good optical properties. In addition, the performance of a ZnFe-LDH photocatalyst can also be enhanced with composites with other agents such as polypyrrole nanofiber [53], diatomite [54], and (BiO)\(_2\)CO\(_3\) [55], etc.

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**References**


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