



TiO₂@ phenyl-functionalized mesoporous silica for removal of bisphenol A from water

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ABSTRACT

A novel phenyl-functionalized mesoporous silica based on TiO₂ supporter was synthesized to remove bisphenol A (BPA) quickly, which is a representative endocrine disrupting chemical. The prepared sorbents were characterized by nitrogen (N₂) adsorption-desorption measurement (BET) and Fourier transform infrared spectroscopy. The effects of the temperature, initial pH value and initial BPA concentration of solution on the adsorption capacity were investigated. And adsorption behaviors of the adsorbent for BPA, such as adsorption kinetics, isotherms, thermodynamics and effect of coexisting pollutants, were also investigated. The result indicated the maximum adsorption capacity of TiO₂-Mip-Ms (22 mg g⁻¹) is the two times higher than that of TiO₂-Nip-Ms (11 mg g⁻¹), and the Langmuir model fits better than Freundlich model. The adsorption kinetic of BPA onto TiO₂-Mip-Ms and TiO₂-Nip-Ms belonged to the pseudo-second-order kinetic model, and adsorption rate constant of TiO₂-Mip-Ms (0.0156 g mg⁻¹ min⁻¹) is near three times than that of TiO₂-Nip-Ms (0.0058 g mg⁻¹ min⁻¹). The BPA adsorption onto TiO₂-Mip-Ms was thermodynamically feasible and spontaneous.

Keywords: Mesoporous silica; Adsorption; Bisphenol A; Surface molecular imprinting; TiO₂

1. Introduction

The environmental pollution of endocrine disrupting chemicals (EDCs) in water sources has drawn considerable scientific and public attention over the past decade [1–3] because they can interfere the reproductive system of wild-life and human even at very low concentration (ppb level). Bisphenol A (2,2-bis(4-hydroxyphenyl)propane; BPA), which is one of these EDCs [4–6], has been widely used in the pro-

duction of polycarbonate plastics, epoxy resins and unsaturated polyester-styrene resins as a monomer. As a result, it was reported that BPA had been widely detected in different water sources, with concentration of 81.77–531.85 ng L⁻¹ in water sources, even in the effluents of wastewater treatment plants [7,8]. Therefore, it is urgent to develop an effective and rapid method to remove BPA from water. Physical adsorption is generally regarded as efficient way to remove EDCs because of low cost, easy operation and no producing secondary harmful substances. Activated carbon is a well-known adsorbent of different contaminants [9]. However, it

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has slow adsorption rate and often suffers from pore blocking and a lack of regenerative ability. So it is important to develop fast adsorbents to remove this type of pollutants from water [10]. The molecularly imprinted functionalized materials are considered as a kind of fast adsorbents [11–13].

Mesoporous silica (Ms) is an ordinary adsorbent for many hydrophilic organic contaminants due to their larger surface area, pore size and its hydrophilic nature. In addition, the pore surface of Ms functionalized by anchoring various hydrophobic organic groups by post-synthesis grafting or co-condensation also can be used to remove hydrophobic pollutants [14]. However, the hydrophobic porous silica synthesized by traditional methods exhibits poor site accessibility and slow mass transfer to the target contaminants [15,16]; surface molecular imprinting technology is a good method to prepare fast adsorbent. Titanium dioxide can be used as a supporter of Ms because of its good chemical stability, low cost and nontoxicity [17]. In the present study, we used nano-TiO₂ as a supporter of Ms to prepare TiO₂-mesoporous silica (denoted as TiO₂-Ms) hybrid materials by sol-gel method. During the preparation process, hydrophobic phenyl groups were introduced into the Ms to provide hydrophobic sites, which can interact with two benzene rings of hydrophobic BPA through a strong π - π interaction. Moreover, we adopted surface molecular imprinting technique to further enhance the specific surface area of hydrophobic Ms onto the surface of TiO₂. The prepared adsorbents were used to adsorb BPA from wastewater. The adsorption characteristics of BPA from aqueous solutions over the synthesized materials including the effect factors (initial concentration, solution pH and temperature, co-pollutants), adsorption kinetics, adsorption isotherms and thermodynamics were investigated.

2. Materials and methods

2.1. Materials and reagents

TiO₂ particles (P25, ca. 80% anatase, 20% rutile; BET area, ca. 50 m² g⁻¹) were purchased from Degussa (Germany). Tetraethoxysilane (TEOS; Chengdu Kelong Chemical Reagent Factory, China, 99%), phenyltriethoxysilane (PhTES; Alfa Aesar, 99%), hexadecyltrimethylammonium bromide (CTAB; Aladdin, 99%), BPA (Aldrich, 99%), phenol and 2-Nitrophenol (2-NP; Aldrich, 99%) and the other chemicals were of analytical purity grade (Chengdu Kelong Chemical Reagent Factory, China) without further purification. Stock solution was prepared by dissolving BPA in methanol followed by gradual dilution with ultrapure water to about 20 mg L⁻¹, in which the methanol concentration was about 0.10% (v/v). HPLC-grade acetonitrile was obtained from Merck. Ultrapure water (18.2 M Ω) used throughout the experiments was obtained from laboratory purification system.

2.2. Synthesis of adsorbents

A novel phenyl group functionalized Ms based on TiO₂ particles (denoted as TiO₂-Mip-Ms) was prepared as follows: 0.6 g of CTAB was dissolved in 28 g of hot distilled water (333 K). Then the solution was cooled down, and 2.0 g of 25% ammonia solution was added to adjust the pH value of solution to 11.8. After that, a weighted amount of TiO₂ (Degussa P25, 1.2 g) was dispersed in the solution under vigorous stir-

ring for 2 min followed by treating with ultrasonication for 30 min. In another beaker, 0.6 g of BPA was dissolved in 8 mL of absolute alcohol, and 2.16 g of TEOS and 1.14 g of PhTES were added into it. Then the above-prepared solution was poured into the titanium dioxide suspension under vigorous stirring for 2 min followed by treating with ultrasonication for 30 min, and the mixture solution was aged for 6 h under stirring at room temperature. The resultant white precipitate was filtered, washed with water and ethanol, and dried at 80°C overnight in vacuum. The template CTAB and BPA molecules were removed by suspending solid product in 50 mL of ethanol and 0.15 mol L⁻¹ HCl and refluxing at 80°C for 24 h for two times. The final powder product was filtered and washed with ethanol and water until BPA could not be detected, and dried at 80°C for overnight in vacuum. The phenyl group functionalized Ms based on TiO₂ as a supporter was denoted as TiO₂-Nip-Ms in a similar way to that of the TiO₂-Mip-Ms only without adding BPA. TiO₂-Ms was obtained when neither BPA nor PhTES was used in the preparation process.

2.3. Materials characterization

In order to analyze the functional groups in the prepared adsorbents, infrared absorption spectra of adsorbents were conducted on a Nicolet iS10 FT-IR spectrometer (Thermo Scientific, Germany) using KBr pellet technique over the wave number range of 4,000–400 cm⁻¹ with a resolution of 2 cm⁻¹. The specific surface area and the pore distribution of adsorbents were measured with TriStar II 3020 surface area measurement instrument (Micromeritics Inc., USA) at 77 K following the BET method.

2.4. Adsorption experiments

2.4.1. Adsorption kinetic study

Adsorption kinetic experiments were performed by batch adsorption method at 298 K on a shaker at 200 rpm. A series of samples were prepared by mixing about 0.0050 g of adsorbents (TiO₂-Mip-Ms, TiO₂-Nip-Ms and TiO₂-Ms) with BPA solution (25.00 mL, 10 mg L⁻¹) in a 25 mL conical flask. Samples were withdrawn at selected time intervals from 0 to 80 min and filtered through a 0.45 μ m filter to remove the particles. The exact concentrations of resultant BPA solution were measured by high-performance liquid chromatography (HPLC).

2.4.2. Adsorption isotherm study

Adsorption isotherm experiments were carried out by mixing approximately 0.005 g of adsorbents (TiO₂-Mip-Ms, TiO₂-Nip-Ms and TiO₂-Ms) with 25.00 mL BPA solution in a 50 mL conical flask. The initial concentration of BPA solution was varied in the range of 10–110 mg L⁻¹. To achieve saturated adsorption, the sample solution was shaken continuously at 200 rpm and 298 K for 1 h and filtered with 0.45 μ m filter. Then concentration of the BPA was analyzed.

2.4.3. Adsorption thermodynamic study

Adsorption thermodynamic experiment was performed by mixing 25.00 mL BPA (10 mg L⁻¹) with about 0.005 g TiO₂-Mip-Ms in a 50 mL conical flask. And the mixture solutions were shaken at different temperature of 293.15, 313.15 and

333.15 K in a temperature-controlled mechanical shaker (Sukun Sky-200B) for 1 h.

2.4.4. pH study

To order to investigate the effect of pH value of BPA solution on the adsorption capacity, the experiments were carried out by mixing 25.00 mL BPA solution (10 mg L^{-1}) with about 0.005 g TiO_2 -Mip-Ms at different pH values of 2, 4, 5, 7, 9 and 12. The initial pH value was adjusted by hydrochloric acid and sodium hydroxide solutions with concentrations of 0.1 mol L^{-1} and measured using pH meter (Shanghai PHS-3B).

2.4.5. The effect of coexisting pollutants on the adsorption capacity for BPA

In order to investigate recognition ability of TiO_2 -Mip-Ms for BPA, we selected the PP and 2-NP as coexisting contaminants, The adsorption experiments for BPA were carried out by mixing about 0.005 g adsorbents with 25.00 mL BPA solution with different concentration (10 mg L^{-1} , 80 mg L^{-1}) in the presence of 10 mg L^{-1} phenol or 2-NP (10 mg L^{-1}) in a 50 mL conical flask. To achieve saturated adsorption, the sample solution was shaken at 200 rpm and 298 K for 1 h. The concentrations of adsorbates in solution were analyzed by HPLC.

2.4.6. Calculation of adsorption capacity of BPA

The adsorption capacity was calculated according to the following equation (Eq. (1)).

$$q_e = \frac{(C_0 - C_e)V}{m} \quad (1)$$

where q_e (mg g^{-1}) is the equilibrium adsorption amount; C_0 (mg L^{-1}) is the initial concentration of BPA in the solution; C_e (mg L^{-1}) is the equilibrium concentration of BPA in the solution, V (L) is the solution volume; and m (g) is the mass of the adsorbent.

2.5. Analytical method

At the end of the adsorption experiments, the solution containing the adsorbents was filtered through a $0.45 \mu\text{m}$ filter to remove the adsorbents. The concentration of BPA, phenol and 2-NP were analyzed by the HPLC (Agilent Technologies 1200series), in which C18 column (Agilent, Eclipse XDB-C18, $4.6 \text{ nm} \times 150 \text{ mm}$, $5 \mu\text{m}$) was applied. The running temperature was 35°C , and the injection volume was $20 \mu\text{L}$. The UV detector (DAD) wavelength was 226 nm for BPA and phenol, and 278 nm for 2-NP. The mobile-phase composition was acetonitrile/water (45/55, v/v) for BPA/phenol binary solution, and acetonitrile/water (35/65, v/v) for BPA/2-NP binary solution with a flow rate of 1 mL min^{-1} .

3. Results and discussion

3.1. Materials characterizations

In order to ascertain the surface area and pore structure of adsorbents, the N_2 adsorption-desorption isotherms measurement was performed. N_2 adsorption-desorption

isotherms for TiO_2 -Ms, TiO_2 -Mip-Ms, TiO_2 -Nip-Ms and P25 at 77 K were shown in Fig. 1(A). The isotherms of TiO_2 -Ms (a) is identified as classical type IV by IUPAC, which is characteristic of mesoporous structure. Whereas, TiO_2 -Mip-Ms (b) and TiO_2 -Nip-Ms (c) showed type I isotherms, which indicated the materials containing microspores. The isotherm of P25 belongs to type III, which is characteristic in a solid happened on weak gas-solid interaction. Corresponding to BJH pore size distribution of TiO_2 -Ms (a), TiO_2 -Mip-Ms (b), TiO_2 -Nip-Ms (c) and P25 (d) were also presented in Fig. 1(B). The maximum pore size of TiO_2 -Ms was about 2.6 nm, whereas the pore sizes of TiO_2 -Mip-Ms (b) and TiO_2 -Nip-Ms (c) were about 2 nm. Table 1 summarizes the physical parameters of the composite materials obtained from the N_2 adsorption-desorption. The TiO_2 -Ms showed higher large surface area of $262 \text{ m}^2 \text{ g}^{-1}$ than P25 of $50 \text{ m}^2 \text{ g}^{-1}$. When phenyl group was introduced into TiO_2 -Ms, the surface areas and total pore volume of TiO_2 -Mip-Ms and TiO_2 -Nip-Ms decreased to 216 and $140 \text{ m}^2 \text{ g}^{-1}$, respectively. This is because the phenyl functional group

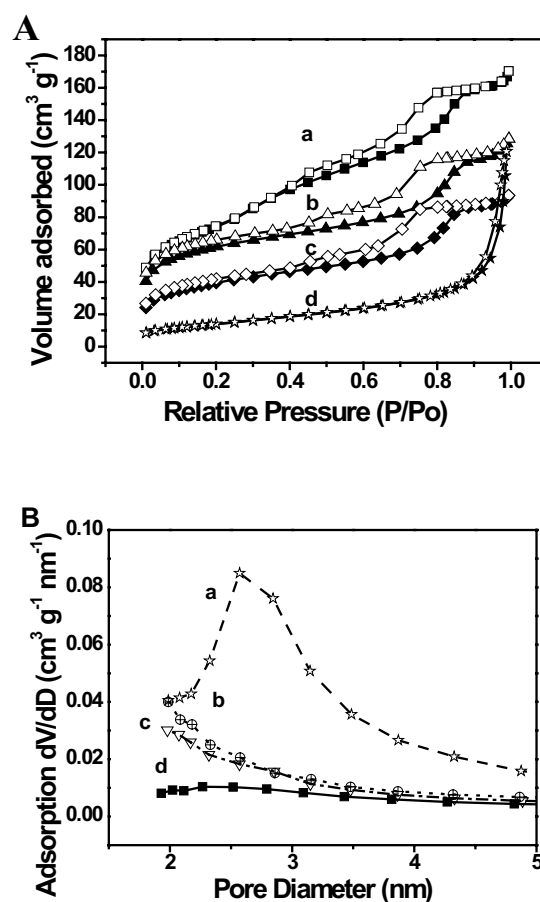


Fig. 1 (A). N_2 adsorption-desorption isotherms and (B) BJH pore size distribution of TiO_2 -Ms (a. TiO_2 /mesoporous silica), TiO_2 -Mip-Ms (b. TiO_2 /phenyl group functionalized imprinting mesoporous silica), TiO_2 -Nip-Ms (c. TiO_2 /phenyl group functionalized non-imprinting mesoporous silica) and P25 (d. open and solid symbols represent desorption and adsorption branches, respectively).

Table 1
BET surface area and pore parameters of TiO₂-Mip-Ms, TiO₂-Nip-Ms, TiO₂-Ms and P25

Adsorbent	Pore diameter (nm)	BET surface area (m ² g ⁻¹)	Total pore volume
TiO ₂ -Mip-Ms	2.0	216	0.15
TiO ₂ -Nip-Ms	2.0	140	0.12
TiO ₂ -Ms	2.6	262	0.25
P25	–	50	–

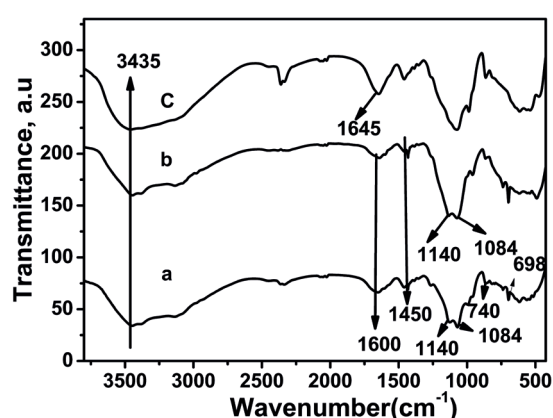


Fig. 2. FT-IR spectra of TiO₂/Nip-MS (a. TiO₂/phenyl group functionalized non-imprinting mesoporous silica) and TiO₂/MS (b. TiO₂/mesoporous silica).

may occupy the internal space of the Ms and made the pores become narrower. The surface area and total pore volume of TiO₂-Mip-Ms were larger than that of TiO₂-Nip-Ms; this phenomenon is attributed to the added BPA during preparation process; BPA may be as a pore expanding agent in this process. In general, the pore diameter should decrease in parallel with the pore volume via modification the organosilanes because the long alkyl chains or other organic groups occupy the internal space of the cavities [18].

To verify the phenyl group has been anchored on the pore surface of Ms successfully. Infrared absorption spectra of TiO₂-Mip-Ms, TiO₂-Nip-Ms and TiO₂-Ms was measured and shown as Fig. 2. It shows that the broad absorption peak of the TiO₂-Mip-Ms, TiO₂-Nip-Ms and TiO₂-Ms near 3435 cm⁻¹ is assigned to the stretching vibration of O-H hydroxyl groups. Si-O-Si vibrational mode appears near 460, 870 and 1072 cm⁻¹. Bending vibration and asymmetrical stretch of Si-OH or Ti-OH is at 1635 cm⁻¹ and 960 cm⁻¹, respectively. Moreover, TiO₂-Mip-Ms and TiO₂-Nip-Ms showed the characteristic absorption peaks of benzene ring at 698 and 740 cm⁻¹ and peak of 1432 cm⁻¹ of C=C stretching vibrations. The peak of 1137 cm⁻¹ indicated the presence of Si-C bond. Fourier transform infrared spectroscopy spectra confirmed that the phenyl group was successfully anchored into the silicate network.

3.2. The effect factors on the adsorption capacity of BPA on the adsorbents

3.2.1. The effect of contact time on the adsorption capacity on the adsorbents

In order to investigate the adsorption rate of BPA on the adsorbents, the adsorption capacity of BPA as a function of contact time was presented in Fig. 3. The results indicated that the adsorption amount of BPA on TiO₂-Mip-Ms increased faster than that of BPA on TiO₂-Nip-Ms from 0 to 40 min, and the BPA adsorption reached the equilibrium in less than 60 min (Fig. 3); However, the adsorption amount of BPA on TiO₂-Ms was very low and did not change at all from 10 to 60 min.

3.2.2. The effect of solution pH on the adsorption capacity of BPA on the adsorbents

In order to investigate the pH influence on the adsorption capacity of BPA on TiO₂-Mip-Ms and TiO₂-Nip-Ms, the experiment was carried out at different solution pH value ranging from 2 to 12. As shown in Fig. 4, the adsorption amount of BPA on adsorbents exhibited a constant with the

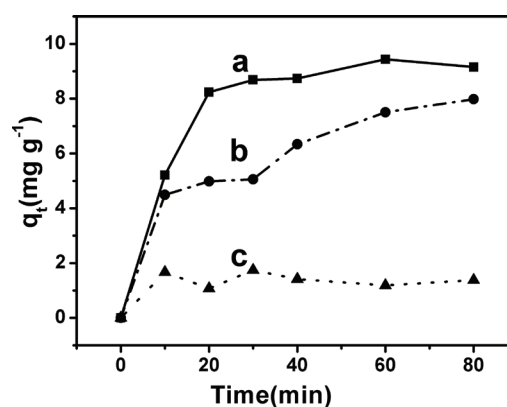


Fig. 3. Adsorption capacities of bisphenol A (BPA) on TiO₂-Mip-Ms (a. TiO₂/phenyl group functionalized imprinting mesoporous silica), TiO₂-Nip-Ms (b. TiO₂/phenyl group functionalized non-imprinting mesoporous silica) and TiO₂-Ms (c. 2 g L⁻¹ dosage and 20 mg L⁻¹ BPA solution) vs. time.

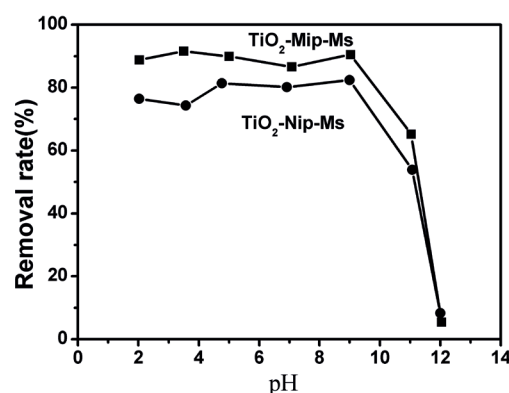


Fig. 4. Effect of solution pH on adsorption capacity of BPA on TiO₂-Mip-Ms (TiO₂/phenyl group functionalized imprinting mesoporous silica) and TiO₂-Nip-Ms (TiO₂/phenyl group functionalized non-imprinting mesoporous silica).

increasing of pH value from 2 to 9, indicating that the binding affinity between BPA and the binding sites did not significantly change under acidic conditions. The adsorption amount of BPA decreased remarkably at a higher pH value ranging from 9 to 11; this phenomenon was observed from other adsorbents, such as activated carbon. The result could be explained by the pKa value of BPA ranging from 9.9 to 10.2, and BPA forms bisphenolate anion, and the electrostatic repulsion between BPA and the sorbent becomes increasingly stronger and overcomes the binding affinity, which led to the adsorption capacity of BPA on the TiO₂-Mip-Ms and TiO₂-Nip-Ms decreased under alkalinity condition.

3.2.3. The effect of initial concentration of BPA on the adsorption capacity

The influence of initial concentration of BPA on the adsorption capacity was investigated with the initial BPA concentration being varied from 10 to 110 mg L⁻¹ as shown in Fig. 5. Among these sorbents, the TiO₂-Ms showed very low adsorption capacities for BPA less than 2 mg g⁻¹, because the TiO₂-Ms has a characteristic of hydrophilic nature compared with the TiO₂-Mip-Ms and TiO₂-Nip-Ms. Whereas, the TiO₂-Mip-Ms and TiO₂-Nip-Ms showed higher adsorption capacity for BPA with maximum values of 22 and 11 mg g⁻¹, respectively, which indicated that phenyl groups introduced into the surface of Ms were hydrophobic and had a strong binding ability for hydrophobic BPA. It was interesting to find that the adsorption capacity of BPA onto TiO₂-Mip-Ms was the two times higher than that of BPA on TiO₂-Nip-Ms. This phenomenon could be assigned to the higher specific area surface of TiO₂-Mip-Ms (216 m² g⁻¹) than that of TiO₂-Nip-Ms (140 m² g⁻¹).

3.2.4. The effect of the temperature of solution on the adsorption capacity

The effect of the temperature on equilibrium adsorption capacity of BPA onto TiO₂-Mip-Ms was studied at 293.15, 313.15 and 333.15 K. The equilibrium adsorption capacity of BPA onto TiO₂-Mip-Ms reduced as the adsorption tempera-

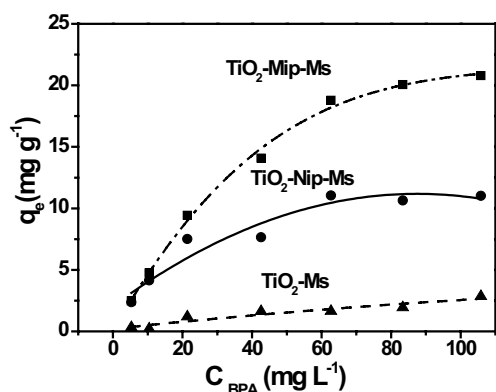


Fig. 5. Effect of initial concentration on adsorption capacity of BPA on TiO₂-Mip-Ms (TiO₂/phenyl group functionalized imprinting mesoporous silica), TiO₂-Nip-Ms (TiO₂/phenyl group functionalized non-imprinting mesoporous silica) and TiO₂-Ms (TiO₂/mesoporous silica).

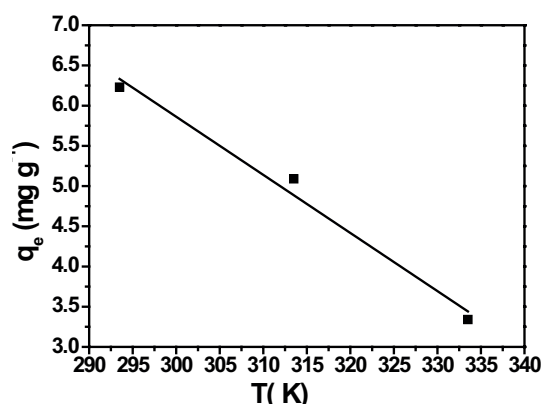


Fig. 6. Effect of temperature of solution on adsorption capacity of BPA on TiO₂-Mip-Ms (TiO₂/phenyl group functionalized imprinting mesoporous silica).

ture increased (Fig. 6), which indicated that the binding of BPA was exothermic in nature.

3.3. Adsorption kinetic study

To investigate adsorption characteristics of BPA, adsorption processes were described by both pseudo-first-order kinetics [19,20] (Eq. (2)) and pseudo-second-order kinetics model [21,22] (Eq. (3)) in this work:

$$\ln \frac{(q_e - q_t)}{q_e} = -k_1 t \quad (2)$$

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t \quad (3)$$

where q_e and q_t (mg g⁻¹) is the adsorption amount of BPA adsorbed at equilibrium and at time t , respectively, and k_1 (1 min⁻¹) and k_2 (g (mg⁻¹·min⁻¹)) is the pseudo-first-order constant and pseudo-second-order constant, respectively.

The experimental data were fitted with the pseudo-first-order kinetics and pseudo-second-order kinetics equations. The results of fitting the experimental data with pseudo-second-order kinetics model on TiO₂-Mip-Ms and TiO₂-Nip-Ms were shown in Fig. 7. The obtained kinetic parameters for BPA onto TiO₂-Mip-Ms and TiO₂-Nip-Ms were summarized in Table 2. The best-fit model was selected according to linear regression correlation coefficient (R^2) and theoretical $q_{e(\text{cal})}$ value. The pseudo-second-order model for BPA adsorption onto TiO₂-Mip-Ms and TiO₂-Nip-Ms can describe better than the Pseudo-first-order model, because the correlation coefficients $R^2 = 0.9920$ and 0.9520 of pseudo-second-order model are higher than 0.9670 and 0.9482 of pseudo-first-order model, respectively. Moreover, the $q_{e(\text{cal})}$ value obtained from pseudo-second-order model fitting was closer to the experimental $q_{e(\text{exp})}$ value than the $q_{e(\text{cal})}$ obtained from pseudo-first-order model. Therefore, the adsorption behaviors of BPA onto TiO₂-Mip-Ms and TiO₂-Nip-Ms belonged to the pseudo-second-order kinetic model. It was also observed that adsorption rate constant of BPA on TiO₂-Mip-Ms (0.0156 g mg⁻¹ min⁻¹) is near three times than that of BPA on TiO₂-Nip-Ms (0.0058 g mg⁻¹ min⁻¹).

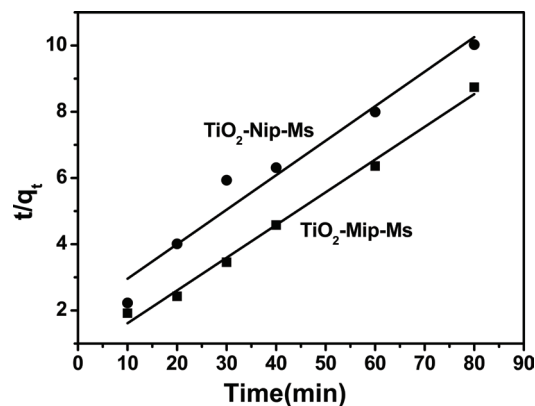


Fig. 7 Pseudo-second-order kinetics plot lines of TiO₂-Mip-Ms (TiO₂/phenyl group functionalized imprinting mesoporous silica) and TiO₂-Nip-Ms (TiO₂/phenyl group functionalized non-imprinting mesoporous silica).

This result can be explained that there are more active sites at the pore surface of TiO₂-Mip-Ms than TiO₂-Nip-Ms.

3.4. Adsorption isotherm study

To understand the adsorbate-adsorbent interactions, we investigated the adsorption isotherms of BPA on the prepared adsorbents. In this work, Langmuir equation (Eq. (4)) and Freundlich equations (Eq. (5)) [23,24] were adopted to simulate the experimental data.

$$\frac{1}{q_e} = \frac{1}{q_m K_L C_e} + \frac{1}{q_m} \quad (4)$$

$$q_e = K_F C_e^{1/n} \quad (5)$$

In Eq. (4), K_L (L mg⁻¹) is the Langmuir constant, which describes the intensity of the adsorption process, and q_m (mg g⁻¹) reflects maximum adsorption capacity. In Eq. (5), K_f is the empirical constant related to bonding energy, which represents the general capacity of BPA adsorbed onto adsorbents for a unit equilibrium concentration. $1/n$, with a range between 0 and 1, reflects adsorption intensity or surface heterogeneity.

The adsorption isotherm parameters were calculated from Langmuir model and Freundlich model and listed in Table 3. The best-fit model was selected according to linear regression correlation coefficient (R^2). It is clear that the Langmuir model fits better than Freundlich model onto TiO₂-Mip-Ms and TiO₂-Nip-Ms, as reflected with correlation coefficients (R^2) of 0.9935 vs. 0.9443, and 0.9808 vs. 0.9206, respectively. The maximum adsorption capacity of TiO₂-Mip-Ms and TiO₂-Nip-Ms for BPA was evaluated to be 21.85 and 11.43 mg g⁻¹, respectively. The higher adsorption capacity of TiO₂-Mip-Ms can be explained that higher specific surface area relative to TiO₂-Nip-Ms. While BPA adsorption on TiO₂-Ms obeyed the Freundlich model, because of higher correlation coefficients (0.9299–0.7731). It indicated that adsorption behavior belonged to homogeneous and monolayer adsorption at the surface of TiO₂-Mip-Ms and TiO₂-Nip-Ms. On the contrary, the adsorption behavior ought to be heterogeneous.

3.5. Adsorption thermodynamics study

The adsorption thermodynamic parameters such as standard free energy change (ΔG°), standard enthalpy change (ΔH°) and standard entropy change (ΔS°) can be estimated with the following equations:

$$\ln k_d = -\frac{\Delta H^\circ}{RT} + \frac{\Delta S^\circ}{R} \quad (6)$$

Table 2
Kinetic parameters for BPA adsorption onto TiO₂-Mip-Ms and TiO₂-Nip-Ms

Adsorbents	Pseudo-first-order kinetics				Pseudo-second-order kinetics		
	κ_1 (min)	$q_{e,cal}$ (mg g ⁻¹)	$q_{e,exp}$ (mg g ⁻¹)	R^2	κ_2 (g mg ⁻¹ min ⁻¹)	$q_{e,cal}$ (mg g ⁻¹)	R^2
TiO ₂ -Mip-Ms	0.069	7.61	9.43	0.9670	0.0156	10.11	0.9920
TiO ₂ -Nip-Ms	0.042	7.36	8.00	0.9482	0.0058	9.59	0.9520

$q_{e,exp}$: adsorption capacity measured from experiment.

$q_{e,cal}$: adsorption capacity estimated by the kinetic model.

R^2 : correlation coefficient.

Table 3
Adsorption isotherm parameters onto TiO₂-Mip-Ms, TiO₂-Nip-Ms and TiO₂-Ms

Adsorbents	Langmuir model			Freundlich model		
	q_m	K_L (L mg ⁻¹)	R^2	K_F	$1/n$	R^2
TiO ₂ -Mip-Ms	21.85	0.2467	0.9935	4.9102	0.3861	0.9443
TiO ₂ -Nip-Ms	11.43	0.2210	0.9808	3.4306	0.2828	0.9206
TiO ₂ -Ms	3.58	0.0203	0.7731	0.1375	0.6425	0.9299

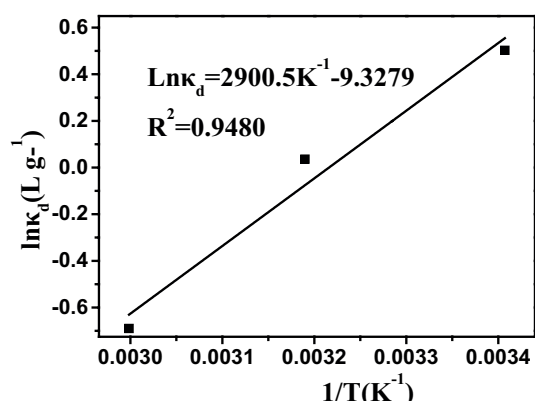


Fig. 8. Linear van't Hoff equation: $\ln \kappa_d$ vs. $1/T$ ($\text{TiO}_2/\text{phenyl}$ group functionalized imprinting mesoporous silica).

$$\Delta G^\circ = \Delta H^\circ - T\Delta S^\circ \quad (7)$$

where R is the universal gas constant, and T (K) is the temperature. Moreover, K_d is the distribution coefficient calculated using the below equation [25]:

$$K_d = \frac{q_d}{C_e} \quad (8)$$

In this work, the distribution coefficient (κ_d) was evaluated in the range of 293.15–333.15 K according to Eq. (8). The van't Hoff equation linearization given by the plot $\ln \kappa_d$ vs. $1/T$ was shown in Fig. 8. From the slope and intercept of the linear fitting, ΔH° and ΔS° were found to be $-24.14 \text{ kJ mol}^{-1}$ and $-77.55 \text{ J mol}^{-1} \text{ K}^{-1}$, respectively. The Gibbs free energy (ΔG°) was calculated from Eq. (7), its value was $-1.38 \text{ kJ mol}^{-1}$ when the adsorption temperature was 293.15 K. It indicated that BPA adsorption onto $\text{TiO}_2\text{-Mip-Ms}$ was thermodynamically feasible and spontaneous.

3.6. The effect of coexisting pollutants on the recognition ability for BPA

In order to assess the effect of introduced phenyl group on the adsorption of BPA in complex matrix and the effect of BPA during preparation, adsorption experiment was carried out in binary solution systems containing 10 mg L^{-1} BPA and PP, or 10 mg L^{-1} BPA and 10 mg L^{-1} 2-NP, respectively. Fig. 9 depicts the equilibrium adsorption amounts of binary component of BPA and PP on $\text{TiO}_2\text{-Mip-Ms}$, $\text{TiO}_2\text{-Nip-Ms}$ and $\text{TiO}_2\text{-Ms}$. The result indicated that the adsorption capacities for BPA onto $\text{TiO}_2\text{-Mip-Ms}$ (8.65 mg g^{-1}) and $\text{TiO}_2\text{-Nip-Ms}$ (8.98 mg g^{-1}) were higher than that of $\text{TiO}_2\text{-Ms}$ (0.35 mg g^{-1}), and the adsorption capacities for PP onto $\text{TiO}_2\text{-Mip-Ms}$ (0.68 mg g^{-1}) and $\text{TiO}_2\text{-Nip-Ms}$ (0.69 mg g^{-1}) were relatively low. This result can be explained that the introduced phenyl group possessed hydrophobic property and had strong interaction with hydrophobic BPA through $\pi\text{-}\pi$ bond. The adsorption capacities for BPA and PP onto $\text{TiO}_2\text{-Ms}$ were very low, because $\text{TiO}_2\text{-Ms}$ possessed hydrophilic nature and hydrogen-bonding is very weak between hydroxyl

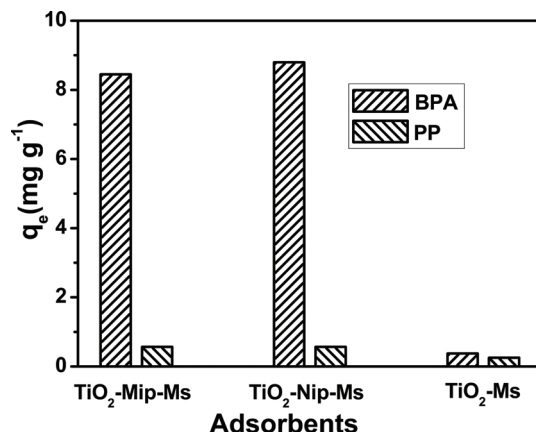


Fig. 9. Equilibrium adsorption capacities in binary component solution containing 10 mg L^{-1} BPA and 10 mg L^{-1} PP on three adsorbents.

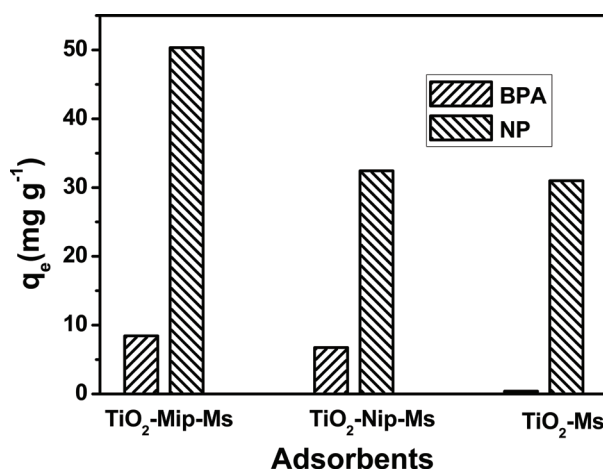


Fig. 10. Equilibrium adsorption capacities in binary component solution containing 10 mg L^{-1} BPA and 10 mg L^{-1} 2-Nitrophenol on three adsorbents.

groups at the surface of Ms and hydroxyl groups of BPA and PP in aqueous solution. We also selected the 2-NP as another co-pollutant because of its similar structure with BPA. The adsorption capacities of the binary component of BPA and 2-NP on the three adsorbents were shown in Fig. 10. We can see that $\text{TiO}_2\text{-Ms}$ has higher adsorption capacities for 2-NP (30.5 mg g^{-1}) because the nitro-group of 2-NP has a stronger hydrogen-binding with $\text{TiO}_2\text{-Ms}$. It is interesting to find that 2-NP adsorption capacities onto $\text{TiO}_2\text{-Mip-Ms}$ (50.0 mg g^{-1}) is nearly two times higher than that on $\text{TiO}_2\text{-Nip-Ms}$ (32.2 mg g^{-1}), and the adsorption capacity for BPA remained unchanged (8.66 mg g^{-1}), which can be explained that 2-NP is somewhat similar to BPA in chemical structure, and it can diffuse easily into large cavities resulting from the BPA pore expanding agent. However, the adsorption capacity of BPA on $\text{TiO}_2\text{-Nip-Ms}$ decreased correspondingly when 2-NP existed, because 2-NP can occupy the surface sites of $\text{TiO}_2\text{-Nip-Ms}$ competitively with BPA.

4. Conclusions

A new type of phenyl-functionalized Ms based on TiO₂ supporter was successfully prepared via surface molecular imprinting technology combined with sol-gel method. The introduction of phenyl group and the existence of much more adsorption sites in Ms layer of TiO₂-Mip-Ms resulted in faster adsorption rate, higher adsorption capacities for BPA than TiO₂-Nip-Ms and TiO₂-Ms. The adsorption behaviors of the adsorbent for BPA belonged to the pseudo-second-order kinetic model and obeyed the Langmuir model. Adsorption process was endothermic and spontaneous.

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